

DATA EXPLORATION AND CPUE STANDARDIZATION FOR THE KOREAN SOUTHERN BLUEFIN TUNA LONGLINE FISHERY (1996-2017)

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ABSTRACT

In this study we standardized southern bluefin tuna, *Thunnus maccoyii* (SBT) CPUE from Korean tuna longline fisheries (1996-2017) using Generalized Linear Models (GLM) with operational data. The data used for the GLMs were catch (number), effort (number of hooks), number of hooks between floats (HBF), fishing location (5° cell), and vessel identifier by year, quarter, and area. We explored CPUE by area and identified two separate areas in which Korean vessels have targeted SBT. SBT CPUE was standardized for each of these areas. We applied two alternative approaches, data selection and cluster analysis, to address concerns about target change through time which can affect CPUE indices. Explanatory variables for the GLM analyses were year, month, vessel identifier, 5° cell, and number of hooks. GLM results for the whole area suggested that location, year, targeting, and month effects were the principal factors affecting the nominal CPUE. The standardized CPUEs for both areas decreased until the mid-2000s and have shown an increasing trend since that time.

INTRODUCTION

Developing indices of abundance using catch per unit effort data requires decisions based on understanding of both the fishery and the population dynamics of the species. This is particularly the case in a multi-species fishery, in which targeting behaviours change seasonally, spatially, and from year to year. Such analyses require careful data exploration, and methods to differentiate fishing practices.

Southern bluefin tuna *Thunnus maccoyii* (SBT) are the target of a high value international fishery, managed by the Commission for the Conservation of Southern Bluefin Tuna (CCSBT). The stock has been assessed as highly depleted, but has recently shown signs of recovery (CCSBT 2017).

Trends in the adult biomass are very influential in determining stock status and catch quotas. The primary index of abundance used to monitor the adult component of the SBT stock (CCSBT 2017) is based on Japanese longline catch and effort data. This analysis of the Korean data was carried out to provide a comparator, based on independently collected and analysed data, with which to validate the trends observed in the Japanese index.

Korean tuna longline fisheries began targeting southern bluefin tuna in the CCSBT convention area in 1991 (Kim et al. 2015). SBT were reported as bycatch before this time, starting in 1972. Catch was initially low but increased to 1,320 mt in 1996, peaked at 1,796 mt in 1998, and thereafter decreased to below

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200 mt in the mid-2000s. In 2008, the catch increased again to 1,134 mt and thereafter fluctuated in a range of 705-1,120 mt due to the national catch limit. The catch in 2017 was 1,080 mt (Fig. 1).

In developing the index, we compare two alternative methods for differentiating targeting practices in the Korean distant water longline data. First, we explore the operational set-by-set data and develop data-based indicators of effort targeting SBT, using the number of hooks between floats (HBF), and the month. Secondly, we use cluster analysis to group the effort into fishing strategies based on the species composition of the catch.

DATA AND METHODS

Set by set catch and effort data were compiled by the Korean National Institute of Fisheries Science. Data were selected with the criterion that when a vessel reported the capture of at least one SBT in a month, all effort for the vessel-month was included.

The CCSBT Korean tuna longline vessels fishing for SBT have mainly operated in two locations to the south of 35°S either between 10°E-50°E (within statistical area 9) or between 90°E-120°E (within statistical area 8) (Figure 2). Effort has focused on western areas (statistical area 9) from March to July/August and shifted to the east (statistical area 8) from July/August until December (Figure 3). In general, there has been more fishing effort in the west.

The fields reported in the operational (set by set) data were catch (number), effort (number of hooks), floats (number of floats), vessel id, location to 1° cell of latitude and longitude, date, and catch in numbers of southern bluefin tuna (SBT), bigeye (BET), yellowfin (YFT), albacore (ALB), swordfish (SWO), black marlin (BLM), blue marlin (BUM), striped marlin (MLS), sailfish (SFA), skipjack (SKJ), sharks (SHA), and other species (OTH).

Dates were converted to months and quarters. Moon phase was used to calculate the relative lunar illumination for each date, using the R package *lunar* (Lazaridis 2014). Spatial positions were classified into 5° cells, and CCSBT statistical areas. The numbers of hooks between floats (HBF) were calculated by dividing hooks by floats and rounding to the nearest whole number.

For CPUE standardization, data were cleaned by removing sets in which there were fewer than 1,000 hooks.

Data were plotted to explore trends in total catch through time; the spatial and seasonal distributions of effort; and patterns in operational characteristics such as HBF and hooks per set. We examined patterns through time and among species in the nominal catch rates by year-quarter and statistical area, and compared them with patterns in the proportions of sets with no catch of each species. We plotted maps of the species composition through time, to identify possible changes in fishing behaviour or population composition.

To further explore changes in the fishery and identify periods of change, we plotted the participation of vessels in the fleet, sorted first by the start date and then by the end date of participation in the fishery.

Several approaches were used to explore changes in effort distribution and concentration through time. For each statistical area and for each year, we plotted the numbers of 5°x5° and 1°x1° cells fished and the

average number of operations per fished cell. We defined two separate core SBT fishing areas: with statistical areas 9 in the west from March-October, and statistical area 8 in the east from July-December.

Indices of fishing effort concentration were also calculated, including the Gini coefficient (Gini 1912) and Gulland's index of concentration (Gulland 1956). The Gini coefficient is best known as an indicator of wealth concentration, but can be used to measure aggregation of any quantity. We use it to estimate the spatial aggregation of the catch of each species, and effort, in each region. A higher Gini coefficient indicates that more of the catch (or effort) is being taken from fewer spatial cells. We estimated values separately for each year, where the values y_i are catches or effort per $5^\circ \times 5^\circ$ cell, ranked from lowest to highest, and including zeroes for unfished cells. Cell areas are assumed to be uniform.

$$Gini = \frac{2 \sum_{i=1}^n i y_i}{n \sum_{i=1}^n y_i} - \frac{n+1}{n}$$

Gulland's index of concentration measures the extent to which a fleet has concentrated its fishing effort in areas with higher than average catch rates (Harley 2009). The weighted version of the index is calculated as follows, where y_i is the catch in the i th stratum, e_i is the effort in the i th stratum, and N is the number of exploited strata.

$$Gulland = \frac{\sum_{i=1}^n y_i}{\sum_{i=1}^n e_i} \cdot \frac{1}{\sum_{i=1}^N \frac{y_i}{e_i N}}$$

This index varies from year to year depending on both the distribution of the effort, and the distribution of the catch rates. If effort is evenly distributed with respect to catch rate then the index will average 1, whereas it will be higher than 1 if effort is preferentially targeted to areas with higher than average catch rate (Hoyle 2014).

Given the spatial and seasonal separation of fishing in these two areas, and potentially different size distributions, we standardized data separately for each area.

Data from the period 1996-2017 were used in CPUE standardizations. Data prior to 1996 were not used in this study as they included insufficient reliable data from vessels targeting SBT.

CPUE standardization methods generally followed the approaches used by Hoyle and Okamoto (2011) and Hoyle et al. (2015), with some modifications. Parts of the methods text below are the same as these articles. R code is also used from examples presented in Hoyle et al. (2014).

Target change

Target change can be a significant problem for CPUE standardization since it can bias CPUE trends. Analyses were carried out using two alternative approaches to address target change. The first approach removed effort considered unlikely to have targeted southern bluefin tuna. The second approach applied cluster analysis of species composition to separate effort into groups that may have used different targeting methods, and then included the categorical cluster variable in the standardization model.

The data selection method firstly removed sets in which HBF was less than 9 or greater than 12. Secondly, data from each statistical area were selected for the periods in which most SBT were caught.

Effort for area 8 was included for the months July to December; and effort for area 9 was included for March to October.

Clustering

In the clustering approach, we clustered all data for areas 8 and 9 using the approach applied by Hoyle et al. (2015). We removed all sets with no catch of any of the species, and then aggregated by vessel-month. Set-level data contains variability in species composition due to the randomness of chance encounters between fishing gear and schools of fish, which leads to some misallocation of sets using different fishing strategies. Aggregating the data reduces variability and misallocation of sets. For these analyses we aggregated the data by vessel-month, assuming that individual vessels tend to follow a consistent fishing strategy through time. One trade-off with aggregation is that vessels may change their fishing strategy within a month, which will result in misallocation of sets. For the purposes of this paper we refer to aggregation by vessel-month as trip-level aggregation, although the time scale is (for distant water vessels) in most cases shorter than a fishing trip.

We calculated proportional species composition by dividing the catch in numbers of each species by catch in numbers of all species in the vessel-month. Thus, the species composition values of each vessel-month summed to 1, ensuring that large catches and small catches were given equivalent weight. The data were transformed by centring and scaling, to reduce the dominance of species with higher average catches. Centring was performed by subtracting the column (species) mean from each column, and scaling was performed by dividing the centred columns by their standard deviations.

We clustered the data using the hierarchical Ward *hclust* method, implemented with function *hclust* in R, option 'Ward.D', after generating a Euclidean dissimilarity structure with function *dist*. This approach differs from the standard Ward D method which can be implemented by either taking the square of the dissimilarity matrix or using method 'ward.D2' (Murtagh and Legendre 2014). However in practice the method gives similar patterns of clusters to other methods, more reliably than 'ward.D2' (Hoyle et al 2015).

Data were also clustered using the *kmeans* method, which minimises the sum of squares from points to the cluster centres, using the algorithm of Hartigan and Wong (1979). It was implemented using function *kmeans* in the R stats package (R Core Team 2016).

Selecting the number of groups

We used several subjective approaches to select the appropriate number of clusters, which in most cases suggested similar numbers of groups. First, we considered the number of major targeting strategies likely to appear in the dataset, based on understanding and exploration of the data. Second, we applied *hclust* to transformed trip-level data and examined the hierarchical trees, subjectively estimating the number of distinct branches. Third, we ran *kmeans* analyses on untransformed trip-level data with number of groups *k* ranging from 2 to 25, and plotted the deviance against *k*. The optimal group number was the lowest value of *k* after which the rate of decline of deviance became slower and smoother. Finally, following Winker et al. (2014) we applied the *nScree* function from the R *nFactors* package (Raiche and

Magis 2010), which uses various approaches (Scree test, Kaiser rule, parallel analysis, optimal coordinates, acceleration factor) to estimate the number of components to retain in an exploratory PCA.

Plotting and data selection

We plotted the hclust clusters to explore the relationships between them and the species composition and other variables, such as HBF, number of hooks, year, and set location. Plots included boxplots of a) proportion of each species in the catch, by cluster; b) the distributions of variables by cluster; and c) maps of the spatial distribution of clusters, one map for each cluster.

GLM analyses

The operational data were standardized using generalized linear models in Microsoft R Open 3.3.2 (R Core Team 2016). Analyses were conducted separately for each of the two core areas, and for each of the two target change methods.

Data were prepared by selecting operational data for vessels that had made at least 100 sets, for years in which there had been at least 100 sets, and for 5° cells in which there had been at least 200 sets. Categories with too few sets provide estimates with high uncertainty and low reliability, so this approach removes a few areas, vessels, and time periods that lack much usable information.

Analyses were carried out using generalized linear models that assumed a lognormal distribution with an added constant. The following model, which we call the lognormal constant GLM, was used:

$$\ln(CPUE_s + k) \sim year + vessid + latlong + \lambda(hooks) + g(month) + h(moon) \quad (1)$$

The constant k , added to allow for modelling sets with zero catches of the species of interest, was 10% of the mean CPUE for all sets. The functions λ , g and h were cubic splines with 10, 4, and 4 degrees of freedom respectively. The number of hooks was included in the model to allow for possible hook saturation or other factors associated with hooks per set. The variable $moon$ was the lunar illumination on the date of the set. The variables $year$, $vessid$, and $latlong$ (5° latitude-longitude cell) were fitted as categorical variables.

For the clustering-based approach, models also included a categorical variable for the cluster.

Models did not include HBF because the 'select' method addresses HBF by only including values in the range 9-12. The 'cluster' method addresses targeting independently of HBF, and in any case only 0.8% of sets included HBF outside the 9-12 range.

Data in the lognormal constant GLMs were 'area-weighted', with the weights of the sets adjusted so that the total weight per year-quarter in each 5° cell would sum to 1. This method was based on the approach identified using simulation by Punsly (1987) and Campbell (2004), that for set j in area i and year-qr t with h_{ijt} hooks, the weighting function that gave the least average bias was: $w_{ijt} = \frac{\log(h_{ijt}+1)}{\sum_{j=1}^n \log(h_{ijt}+1)}$. Given

the relatively low variation in number of hooks between sets in a stratum, we simplified this to $w_{ijt} = \frac{h_{ijt}}{\sum_{j=1}^n h_{ijt}}$.

Model fits were examined by plotting the residual densities and using Q-Q plots.

The effects of covariates were examined in exploratory analyses by using the package *influ* (Bentley et al. 2011) to show the influence of each covariate.

Changes in catchability through time were investigated by fitting to the operational data both with and without a term for individual vessel. The two models were designated respectively the 'base model' and the 'vessel-effects model'. Abundance indices were calculated for each model, and normalized to average 1. The indices estimated for each year-quarter were compared by dividing the base model by the vessel effects model, plotting the time series of ratios, and fitting a log-linear regression. The slope of trend in the ratios represented the average annual rate of change in fishing power due to vessel turnover; i.e. the introduction of new vessels and retirement of old vessels. Gradients are shown on the figures, together with confidence intervals.

Indices of abundance were obtained by running the lognormal constant GLM model with the standard settings, including vessel effects. Time effects were obtained by predicting the expected catch rate for each year, for (across all years) the vessel, month and cell with the most sets, lunar illumination of 0.5, and the median number of hooks. The uncertainty associated with the year effect was used as the measure of uncertainty. Indices were normalised by dividing through by the mean of the year effects, giving relative CPUE with mean of 1.

RESULTS

Data exploration

Almost all effort used between 9 and 12 hooks per float (HBF) (Figure 4), while the majority of HBF outside this range came from north of 35S, outside the main SBT targeting area. The number of hooks per set averaged less than 3000 in the period from 1990-95, but since that time has been relatively consistent, averaging a little over 3000.

Mean catch rates by species in the southern statistical areas 7, 8, and 9 are highest for southern bluefin tuna until the mid-2000s. After this time in area 8 and 9 SBT catch rates decrease and other species, particularly albacore, increase. However, in the most recent years the SBT catch rates are again higher than other species (Figure 6). Similarly, the proportion of sets reported with zero SBT catches was low through most of the time series in the southern areas 7 to 9 (Figure 5), but area 9 shows an increase in the proportion of zeroes from 2004 to 2010. Most sets reported no yellowfin catch, and the same applied to bigeye and albacore. However, in area 9 there was a decrease in the proportions of zero albacore catches from about 2000-2010, followed by an increase in the late 2000's. There may have been some albacore targeting in area 9 during this period.

In the northern statistical areas 13 and 1, the tropical bigeye and yellowfin tunas dominate with the highest catch rates, along with albacore (Figure 6). Southern bluefin tuna catch rates are low throughout

the time series, even though data were only selected for vessels reporting at least one SBT in the month. The mean catch rates across all vessels would be lower. The existence of zero SBT catch rates is likely due to vessels being included due to reporting SBT catch during the month in a different statistical area, though some may be due to effort with SBT catch being removed during the cleaning process. Most sets in these areas catch no SBT (Figure 7), and there are few sets with zero catches of bigeye or yellowfin, while intermediate numbers of sets report no albacore catch. Given the low rate of SBT capture in the northern areas, misreporting, species misidentification, and data errors may be a concern, and could explain a sizable proportion of the observed catches.

Statistical areas 14 and 2 in the Indian Ocean are at temperate latitudes between 20S and 35S. Highest catch rates are for yellowfin and (more recently) albacore in the western area 14, and bigeye and albacore in eastern area 2. Since the mid-2000s albacore catch rates have increased markedly and particularly in area 2, suggesting a trend towards targeting this species. Catch rates of SBT have been relatively low throughout the period, consistent with a high proportion of zero SBT sets, suggesting little or no deliberate targeting of SBT by the Korean fleet in these statistical areas.

We mapped the species composition of catch (proportion of SBT in the catch of all species) south of 30S by 5-year period (Figure 8). The proportion of SBT in the catch was high in all periods, increasing further south, but declined steadily in all areas after 2000. This partly reflects targeting of other species, but also reflects increased reporting of sharks and other species. In the post-2010 period there is little SBT taken in statistical area 8 north of about 37S, whereas a high proportion of the catch in this area is albacore (Figure 9). The spatially differentiated targeting in area 8 also has a temporal aspect, with albacore targeting April-July, and SBT targeting July-December. In area 9 to the west, there is less spatial or temporal separation of SBT and ALB catch, with both species caught in the months March-October.

The relationship between the catch per set of southern bluefin tuna and the average catch per set of other major species bigeye, yellowfin, and albacore tuna was quite varied among latitudes and periods (Figure 10). In many cases there appeared to be a positive correlation, but negative correlation in other cases. The relationships do not appear to be simple or consistent across all locations and periods.

Only 39 Korean vessels have participated in the area 8 and 9 fishery since 1996 (Figure 11), with over half of the total reporting their first participation before 2000. New vessels have arrived slowly but regularly. Eight vessels stopped participating in 2009, and five more have stopped since then, but seven others have joined the fishery in that time.

The total number of major (5° x 5° x month) cells fished has varied from year to year (Figure 12) but declined steadily and considerably since the peak in 2008. Over the same period, effort has become more concentrated with more operations per cell. This increasing concentration is also apparent at the minor (1° x 1° x month) cell level (Figure 13). The distribution of effort within major cells was more stable until recently, with similar numbers of minor cells per major cell on average, but in 2017 increased to the highest level yet seen. In the period since 2008 the timing of effort in areas 8 and 9 has changed, gradually moving earlier in the year, though with different timing peaks in each area.

Gulland's index of concentration indicates whether effort is concentrated in areas of high or low catch rate for a species, but estimates can be variable and uncertain where sample sizes are small. Plots for SBT in areas 8 and 9 suggest that effort is generally higher in areas with higher SBT catch rate, since most

points are above 1 (Figure 14). The results for bigeye and albacore are considerably more variable, reflecting the lower catch rates.

Gini coefficients are widely used in many fields to measure the distribution of quantity – with uniform to very uneven distributions represented by low to high Gini coefficients. Estimates for regions 8 and 9 for SBT, bigeye and albacore tuna, and for effort, show similar patterns, with increasing concentration through time (Figure 15).

Target change

The data selection approach aimed to identify effort targeted mostly at southern bluefin tuna, by selecting area 9 data from March-October and area 8 data from July-December. This approach appeared to give reasonable results, but was not entirely successful, as indicated by the high proportions of zero catches in area 9 between 2004 and 2010 (Figure 5).

Applying Ward's D hierarchical cluster analysis at the vessel-month identified strong separation among 2 to 3 groups in statistical areas 8 and 9 (Figure 16). We chose to use three clusters in each area. We preferred to use more clusters where there was uncertainty because unresolved target change can cause bias in indices.

In area 9, clusters 1, 2, and 3 were more strongly represented in the early, middle, and later parts of the time series respectively (Figure 17). Clusters 2 and 3 occur mostly in the period before August, while cluster 1 extends into October. Cluster 3 also has slightly more hooks between floats. Mean number of hooks is higher in cluster 3 and lower in cluster 1. Cluster 2 dominates the southeast of area 9, while cluster 1 dominates the northeast, and cluster 3 the southwest (Figure 18). The species composition of cluster 2 comprises almost entirely southern bluefin tuna, with small amounts of albacore and bigeye tuna (Figure 19). Cluster 1 has the some bluefin tuna along with albacore, and also some bigeye and yellowfin tuna. Cluster 3 includes significant southern bluefin tuna along with similar amounts of 'other' species, along with some sharks and albacore tuna.

In area 8, cluster 1 dominates the early part of the time series, with clusters 2 and 3 more apparent after 2005 (Figure 17). Cluster 1 averages fewer hooks between floats, while the hooks per set are similar for all clusters. Clusters 1 and 2 occur mostly in the second half of the year, while cluster 3 is also represented during March to June. Cluster 1 is well represented across most of the fished area, while cluster 2 occurs at middle latitudes from about 38-43S. Cluster 3 occurs almost entirely in the far north of the area (Figure 18). Cluster 1 is dominated by southern bluefin tuna, with little reporting of other species (Figure 19). Cluster 2 reports similar amounts of southern bluefin tuna, sharks, and 'other' species. Cluster 3 reports more albacore than southern bluefin tuna, with 'other' species, yellowfin and bigeye tuna also reported.

CPUE standardization

The models were initially applied to data from the first target change method, using the data selection process. Table 1 shows the results of dropping each variable from the lognormal constant GLMs, indicating that all explanatory variables were statistically significant, with the year, location, and month effects the largest factors affecting the nominal CPUE. It is common in CPUE standardizations for all variables to be statistically significant. However, lack of independence is to be expected in observational fisheries data and tends to result in overestimation of statistical significance.

Patterns were broadly similar for both approaches to addressing target change (Figure 20). There were differences in the 2004-2006 period in area 9, with standardized catch rates higher for the clustered analysis. The clustering approach may be accounting better for the apparent switch towards targeting albacore during this time. The presence of more zero SBT catches in area 9 from 2004-2010 suggests that the data during that period may include more effort targeted at other species. Such 'contamination' of the effort would tend to bias the indices low, and this is a period when the indices are the lowest in the time series. It is therefore useful to separate sets with different fishing strategies. Applying cluster analysis to separate may be the best approach for this period.

Comparison of standardized and unstandardized CPUE series shows them to be quite similar. The largest change is for the area 8 indices in the most recent years, where the 'selected data' standardized indices first dips below the nominal and then climbs above it. The 'clustered data' indices are more stable overall, but also increase strongly in 2016.

The influence plots (Figures 21-25) showed the patterns of the parameter estimates at the top of each plot, and the influence of each parameter on the year effect on the right side of each plot. Note that the influence scales (bottom right) differ among plots. The influences of all variables are summarised in Figure 27. To save space only the clustering plots are presented.

Vessel effects (Figure 21) were quite variable, with a few vessels having significantly lower SBT catch rates. The influence of the vessel effects changed through the time series, with the low number of vessels causing significant variability.

Spatial effects (Figure 22) showed significant variation in catch rates, with more variation in area 9 than area 8. In area 9 there was an initial trend towards fishing in areas with lower average catch rates. It would be useful to explore whether the areas of highest catch rate have moved through time. This would be difficult to determine from Korean data alone, since fishing activity is currently very concentrated spatially. Given the behaviour of the species, areas of highest catch rate are also likely to move during the year, and we do not account for this in the model. This is likely to increase uncertainty in the indices.

The effects of the number of hooks per set on catch rates (Figure 23) were difficult to interpret. In eastern area 8 there were relatively small differences by hook number across the range of data with most hooks, and minimal influence on year effects, apart from 2016 when effort was low and localised with only one vessel fishing. In area 9 there were larger differences, and apparent influence on the year effects, with catchability averaging about 5% above the mean in 2010-16. Sets with more than about 3150 hooks tended to catch more SBT than sets with fewer hooks. This may reflect a mixture of targeting methods in area 9, with different fishing methods using different numbers of hooks. In area 9 there were more sets with fewer hooks between 2004 and 2007, a period during which there were more zero SBT sets than at most other times.

The effect of month was strong in both the eastern and western areas (Figure 24). In both areas, the highest catch rates were obtained in July and August. The seasonality of fishing effort changed through time, with the model suggesting that mean catchability in area 8 was over 10% higher than the average in 2014-15, and almost 10% below average in area 9 in 2011-15.

Catch rates appeared to vary moderately with lunar illumination (Figure 25). Longline catch rates of other pelagic fish such as bigeye tuna are known to be affected by moon phase (Poisson et al. 2010). The

patterns we observed differed between the two areas and may be artefacts of lack of independence in the data. Fishing effort is distributed relatively evenly across all phases of the moon, so moon phase has almost no influence on the year effects.

The distribution of the cluster variable *hcltrp* changed considerably though time, as the behaviour of the fleet changed with the abundance of the target species (Figure 26). There were also relatively large differences in catch rate between the clusters, so this variable was quite influential on the indices.

The summary plots showing the effects of all the influence plots shows how many factors affect the index trends (Figure 27). The variables have mostly similar effects in the 'selected data' and 'clustered data' analyses.

Like the abundance indices, the influence estimates are conditional on the model, which assumes that there are no interactions between the different effects. However, interactions may be expected, such as variation between years in the timing and location of higher catch rates, due to environmental variation affecting tuna movements. The small sample sizes limit the ability to model interaction terms, but there may be value in exploring space-time interactions using smoothing splines in *mgcv*.

Diagnostic frequency distributions and QQ-plots (Figure 28) suggest that the data fitted the GLM adequately.

Patterns in the indices (Figure 29) differ somewhat between area 8 and 9. Both sets of indices decreased until the mid-2000s, and subsequently increased, particularly in the last few years. Lack of data prevents the estimation for eastern area 8 from 2003-2007 (Table 2). The recent effort in region 8 is so low and concentrated that the estimates are likely to be highly variable, and not very reliable. There was no effort and no index for region 8 in 2017. Effort in region 9 is also becoming more concentrated, so indices are also likely to be getting less reliable.

The ratios of analyses with and without vessel effects did not suggest significant trends in fishing power. Confidence limits on the trend estimates were wide relative to the potential effects, as expected since sample sizes and the numbers of vessels involved were small.

Discussion

This application of cluster analysis to the data used in the southern bluefin tuna model suggests that the approach may be more useful than the alternative approach based on data selection. The cluster analysis approach identified patterns that are consistent with our understanding of the fishery, and it changed the indices during the 2004-2006 period in a way that seems likely to better track the abundance. Other changes to the indices cannot be characterised as improvements or otherwise, given the sparse data and uncertainty in all estimates. We prefer the 'clustered data' indices, but either set of indices might be considered.

Estimates of change in fishing power were inconclusive. Trends in fishing power estimated this way represent the effects of changes in the fleet composition, and do not account for effects where vessels change their equipment or their fishing behaviour.

A striking pattern emerging in the fishery is that the Korean fleet has greatly concentrated its effort during the last 5 years. As catch rates have increased, the fleet has significantly reduced the area fished to catch its quota, and in 2017 was more concentrated than ever before. One effect of this concentration

was that in 2017 there was no effort in the east in statistical area 8, resulting in no update of the index for that area. Another is that the Gulland index of concentration has become difficult to calculate and essentially lost its usefulness as the fleet has concentrated.

CPUE indices are associated with the selectivity of the fishery that provides the catch and effort data. Improving the selectivity estimates for the fishery will provide better estimates of stock size and stock status.

REFERENCES

- Bentley, N., Kendrick, T.H., Starr, P.J., Breen, P.A. (2011) Influence plots and metrics: tools for better understanding fisheries catch-per-unit-effort standardizations. *Ices Journal of Marine Science*, 69(1): 84-88. 10.1093/icesjms/fsr174
- Campbell, R.A. (2004) CPUE standardisation and the construction of indices of stock abundance in a spatially varying fishery using general linear models. *Fisheries Research*, 70(2-3): 209-227.
- CCSBT (2017) Report of the Twenty Second Meeting of the Scientific Committee.
- Gini, C. (1912) Variabilità e mutabilità. *Reprinted in Memorie di metodologica statistica (Ed. Pizetti E, Salvemini, T). Rome: Libreria Eredi Virgilio Veschi, 1.*
- Glazer, J., Butterworth, D. (2002) GLM-based standardization of the catch per unit effort series for South African west coast hake, focusing on adjustments for targeting other species. *African Journal of Marine Science*, 24: 323-339.
- Gulland, J.A. (1956) On the fishing effort in English demersal fisheries. *Fisheries Investigations Series II*, 20(5): 41 pp.
- Harley, S. (2009) Spatial distribution measures for the analysis of longline catch and effort data: 17.
- Hartigan, J.A., Wong, M.A. (1979) Algorithm AS 136: A k-means clustering algorithm. *Journal of the Royal Statistical Society. Series C (Applied Statistics)*, 28(1): 100-108.
- Helidoniotis, F. (2016) An updated CPUE Index based on a GAMM. CCSBT-CPUE/1606/08, CCSBT-CPUE/1606/08.
- Hoyle, S.D. (2014) Spatial considerations in bigeye and yellowfin CPUE from Japanese and Taiwan, China longline fisheries in the Indian Ocean, IOTC-2014-WPTT16-27 Rev_1.
- Hoyle, S.D., Assan, C., Chang, S.-T., Fu, D., Govinden, R., Kim, D.N., Lee, S.I., Lucas, J., Matsumoto, T., Satoh, K., Yeh, Y.-m., Kitakado, T. (2017) Collaborative study of tropical tuna CPUE from multiple Indian Ocean longline fleets in 2017. IOTC-2017-WPTT19-32. *Indian Ocean Tuna Commission, Working Party on Tropical Tunas.*
- Hoyle, S.D., Langley, A.D., Campbell, R.A. (2014) Recommended approaches for standardizing CPUE data from pelagic fisheries.

Hoyle, S.D., Okamoto, H. (2011) Analyses of Japanese longline operational catch and effort for bigeye and yellowfin tuna in the WCPO, WCPFC-SC7-SA-IP-01. *Western and Central Pacific Fisheries Commission, 7th Scientific Committee*, Pohnpei, Federated States of Micronesia, 2011.

Hoyle, S.D., Okamoto, H., Yeh, Y.-m., Kim, Z.G., Lee, S.I., Sharma, R. (2015) IOTC-CPUEWS02 2015: Report of the 2nd CPUE Workshop on Longline Fisheries, 30 April – 2 May 2015, IOTC-2015-CPUEWS02-R[E]: 126.

Itoh, T. (2017) Change in operation pattern of Japanese southern bluefin tuna longliners in the 2016 fishing season. CCSBT-CPUE/1706/09, CCSBT-CPUE/1706/09: 18.

Itoh, T., Takahashi, N. (2017) Update of the core vessel data and CPUE for southern bluefin tuna in 2017. CCSBT-CPUE/1706/08, CCSBT-CPUE/1706/08: 19.

Itoh, T., Takahashi, N. (2018) Documentation on the method for preparation of CPUE data by Japan. *CCSBT Data CD Documentation*.

Kim, Z.G., Kim, D.N., Lee, S.I., Kwon, Y., Cha, H.K. (2015) CCSBT-ESC/1509/SBT, 2015 Annual National Report of Korean SBT Fishery. *20th Extended Scientific Committee of the CCSBT*, Incheon, Republic of Korea, 1 - 5 Sep 2015.

Lazaridis, E. (2014) lunar: Lunar Phase & Distance, Seasons and Other Environmental Factors (Version 0.1-04). Available from <http://statistics.lazaridis.eu>. <http://statistics.lazaridis.eu>

Maunder, M.N., Punt, A.E. (2004) Standardizing catch and effort data: a review of recent approaches. *Fisheries Research*, 70(2-3): 141-159. 10.1016/j.fishres.2004.08.002

McKechnie, S., Tremblay-Boyer, L., Pilling, P. (2017) Background analyses for the 2017 stock assessments of bigeye and yellowfin tuna in the western and central Pacific Ocean, WCPFC-SC13-2017/SA-IP-06. *WCPFC Scientific Committee 13*, Rarotonga, Cook Islands, 9–17 August 2017.

Murtagh, F., Legendre, P. (2014) Ward's hierarchical agglomerative clustering method: which algorithms implement Ward's criterion? *Journal of Classification*, 31(3): 274-295.

Poisson, F., Gaertner, J.-C., Taquet, M., Durbec, J.-P., Bigelow, K. (2010) Effects of lunar cycle and fishing operations on longline-caught pelagic fish: fishing performance, capture time, and survival of fish. *Fishery Bulletin*, 108: 268-281.

Punsly, R. (1987) *Estimation of the relative annual abundance of yellowfin tuna, Thunnus albacares, in the eastern Pacific Ocean during 1970-1985*. I-ATTC, LA JOLLA, CA ().

R Core Team (2016) R: A Language and Environment for Statistical Computing. R Foundation for Statistical Computing, Vienna, Austria. <https://www.R-project.org/>

Raiche, G., Magis, D. (2010) nFactors: Parallel analysis and non graphical solutions to the Cattell Scree Test. *R package version*, 2(3).

Thorson, J.T. (2014) Standardizing compositional data for stock assessment. *Ices Journal of Marine Science*, 71(5): 1117-1128.

Thorson, J.T., Barnett, L.A. (2017) Comparing estimates of abundance trends and distribution shifts using single-and multispecies models of fishes and biogenic habitat. *Ices Journal of Marine Science*, 74(5): 1311-1321.

Thorson, J.T., Fonner, R., Haltuch, M.A., Ono, K., Winker, H. (2016) Accounting for spatiotemporal variation and fisher targeting when estimating abundance from multispecies fishery data. *Canadian Journal of Fisheries and Aquatic Sciences*, 73(999): 1-14.

Thorson, J.T., Haltuch, M.A. (2018) Spatio-temporal analysis of compositional data: increased precision and improved workflow using model-based inputs to stock assessment. *Canadian Journal of Fisheries and Aquatic Sciences*(ja).

Winker, H., Kerwath, S.E., Attwood, C.G. (2014) Proof of concept for a novel procedure to standardize multispecies catch and effort data. *Fisheries Research*, 155: 149-159.

TABLES

Table 1: Degrees of freedom, Deviance, and delta AIC results from lognormal (CPUE + k) GLMs for statistical areas 8 and 9.

	Stat area 9			Stat area 8		
	Df	Deviance	ΔAIC	Df	Deviance	ΔAIC
<none>		148.0	0		36.4	0
Year	20	174.6	2413	14	42.3	1285
Latlong	18	160.1	1129	10	38.1	365
ns(hooks, 10)	10	150.1	197	5	36.7	55
Vessid	28	159.0	1003	21	37.4	182
ns(month, df = 4)	4	156.7	843	3	37.6	266
ns(moon, df = 4)	4	148.2	10	4	37.3	185

Table 2: Lognormal constant indices for statistical areas 9 and 8, for selected data (left) and clustered data (right).

Year	Selected data				Clustered			
	Stat area 9	CV	Stat area 8	CV	Stat area 9	CV	Stat area 8	CV
1996	1.28	0.04	0.96	0.04	1.23	0.04	1.20	0.04
1997	0.94	0.03	0.63	0.02	1.08	0.03	0.77	0.03
1998	0.90	0.03	0.71	0.02	0.97	0.03	0.84	0.03
1999	1.00	0.03	0.59	0.02	0.99	0.03	0.74	0.03
2000	0.83	0.03	0.61	0.02	0.85	0.04	0.76	0.03
2001	0.99	0.03	0.69	0.03	0.96	0.03	0.78	0.03
2002	0.93	0.03	0.39	0.02	0.93	0.03	0.50	0.03
2003	0.73	0.03	-	-	0.81	0.04	-	-
2004	0.32	0.03	-	-	0.42	0.03	-	-
2005	0.10	0.05	-	-	0.23	0.05	-	-
2006	0.56	0.04	-	-	0.68	0.04	-	-
2007	0.43	0.03	-	-	0.54	0.03	-	-
2008	0.98	0.03	1.00	0.02	0.99	0.03	1.08	0.02
2009	0.73	0.03	0.65	0.03	0.80	0.03	0.68	0.03
2010	0.77	0.03	0.73	0.02	0.81	0.03	0.75	0.03
2011	2.16	0.05	0.98	0.02	1.88	0.05	1.02	0.03
2012	1.68	0.03	1.19	0.03	1.54	0.04	1.25	0.04
2013	1.35	0.05	1.60	0.04	1.28	0.05	1.17	0.05
2014	2.37	0.05	1.84	0.05	2.05	0.05	1.02	0.05
2015	1.25	0.05	1.05	0.04	1.13	0.06	1.11	0.05
2016	1.76	0.04	2.82	0.05	1.56	0.04	2.56	0.06
2017	1.87	0.05			1.63	0.05		

FIGURES

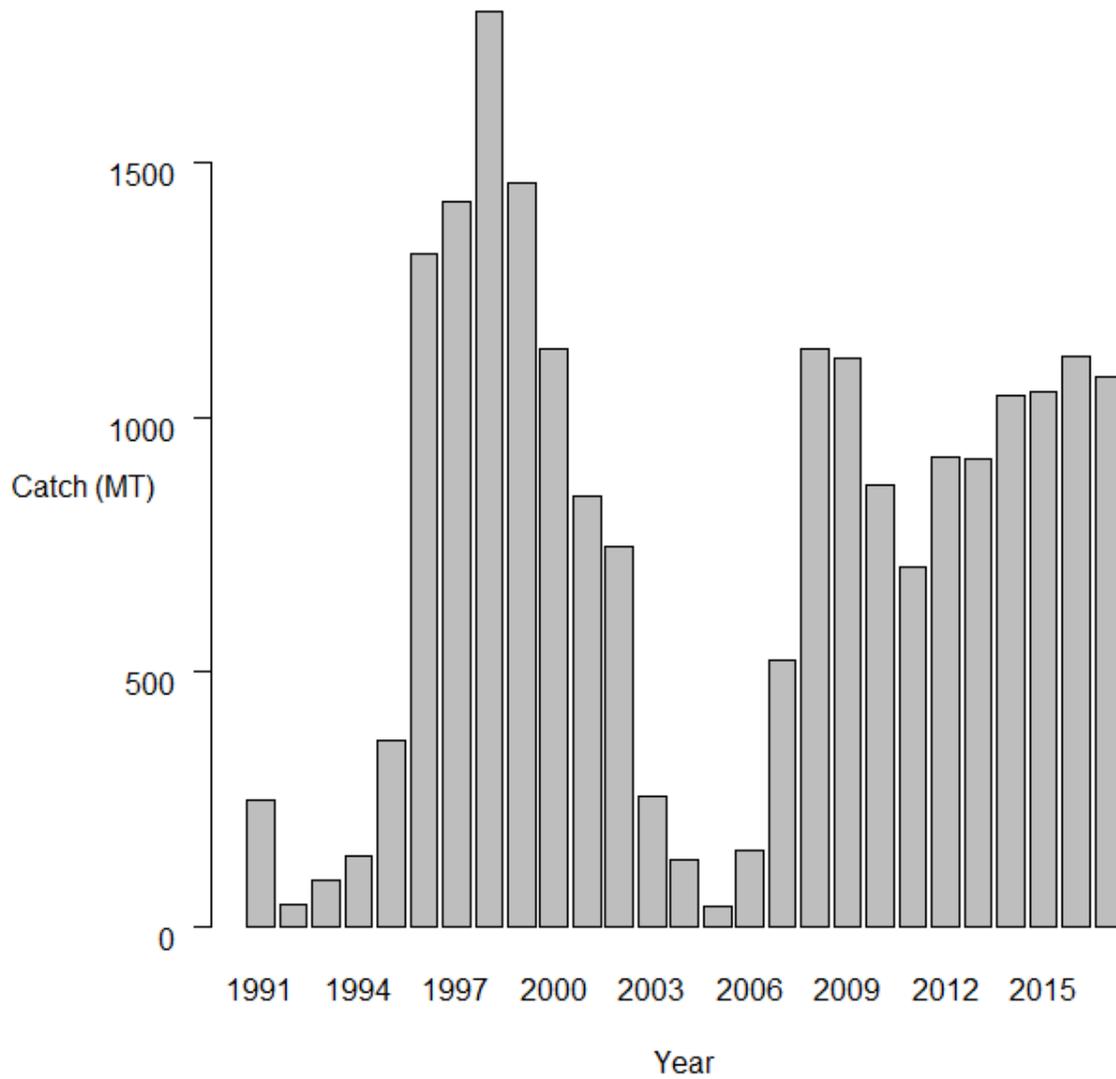


Figure 1: The annual Korean SBT catches in the CCSBT convention area, 1991 - 2017.

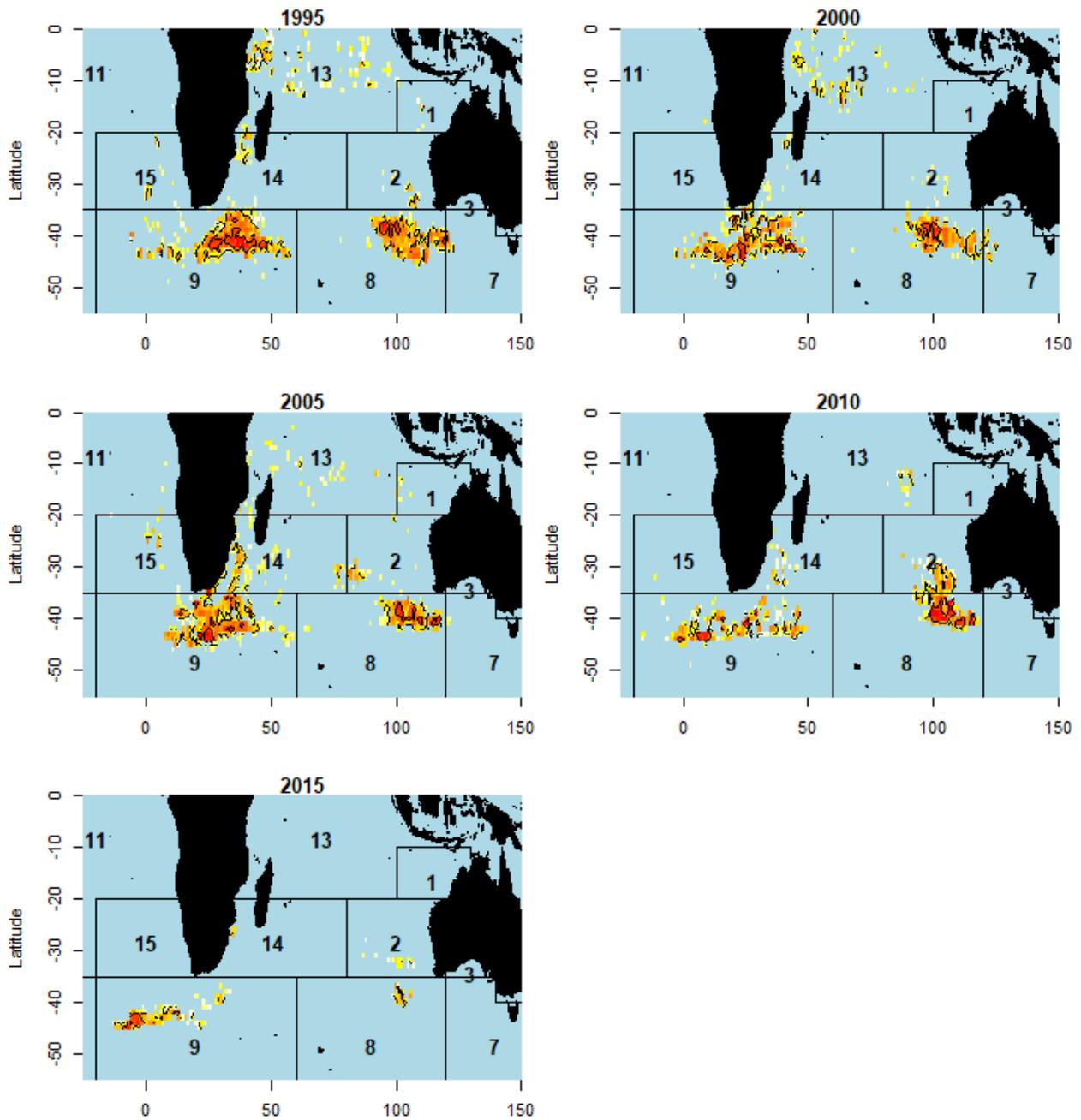


Figure 2: Map showing the core areas of Korean tuna longline vessels fishing for SBT, aggregated by 5-year period. Red colour indicates higher fishing effort, n numbers of hooks.

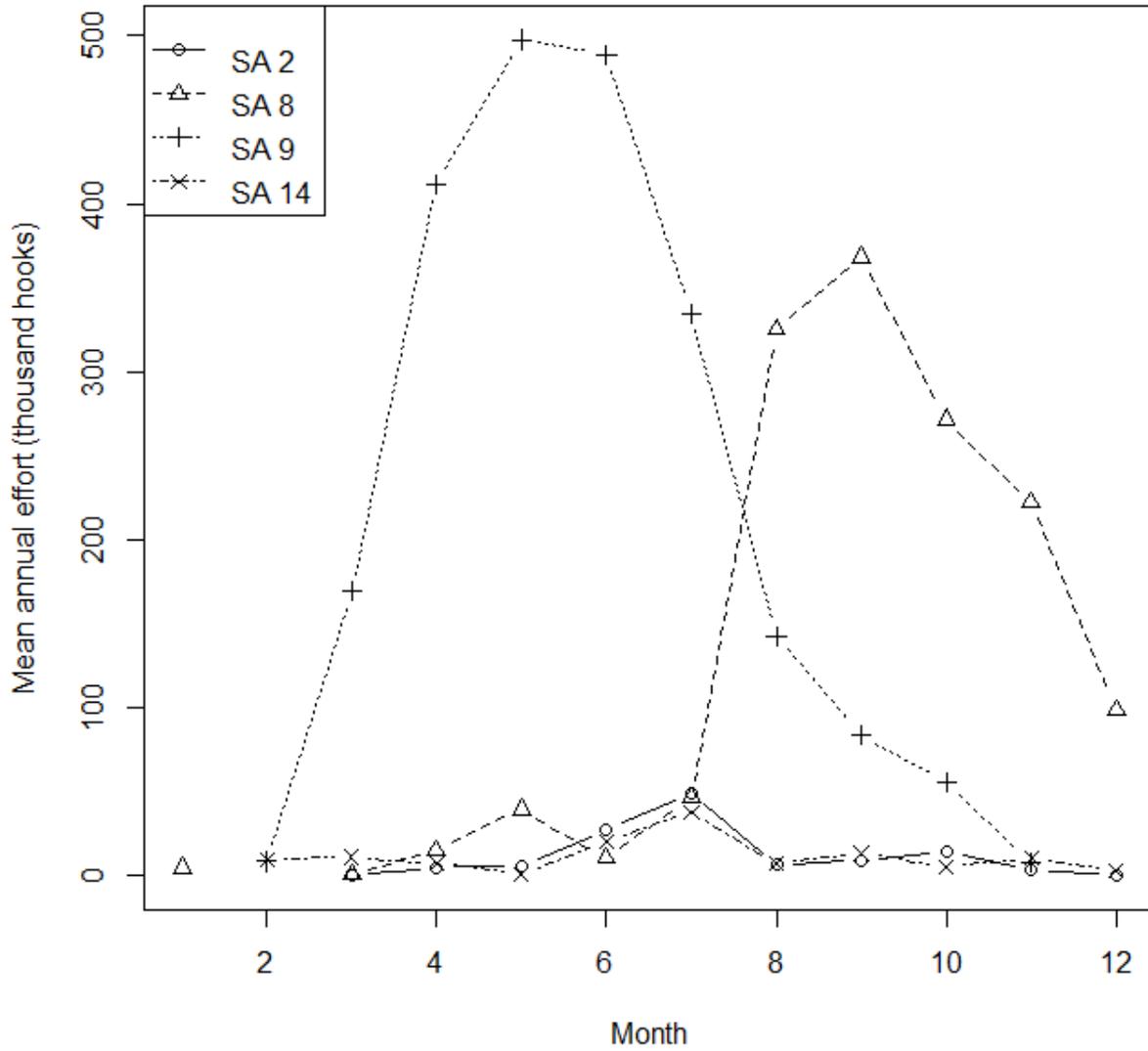


Figure 3: Mean annual effort in thousands of hooks, by month and statistical area.

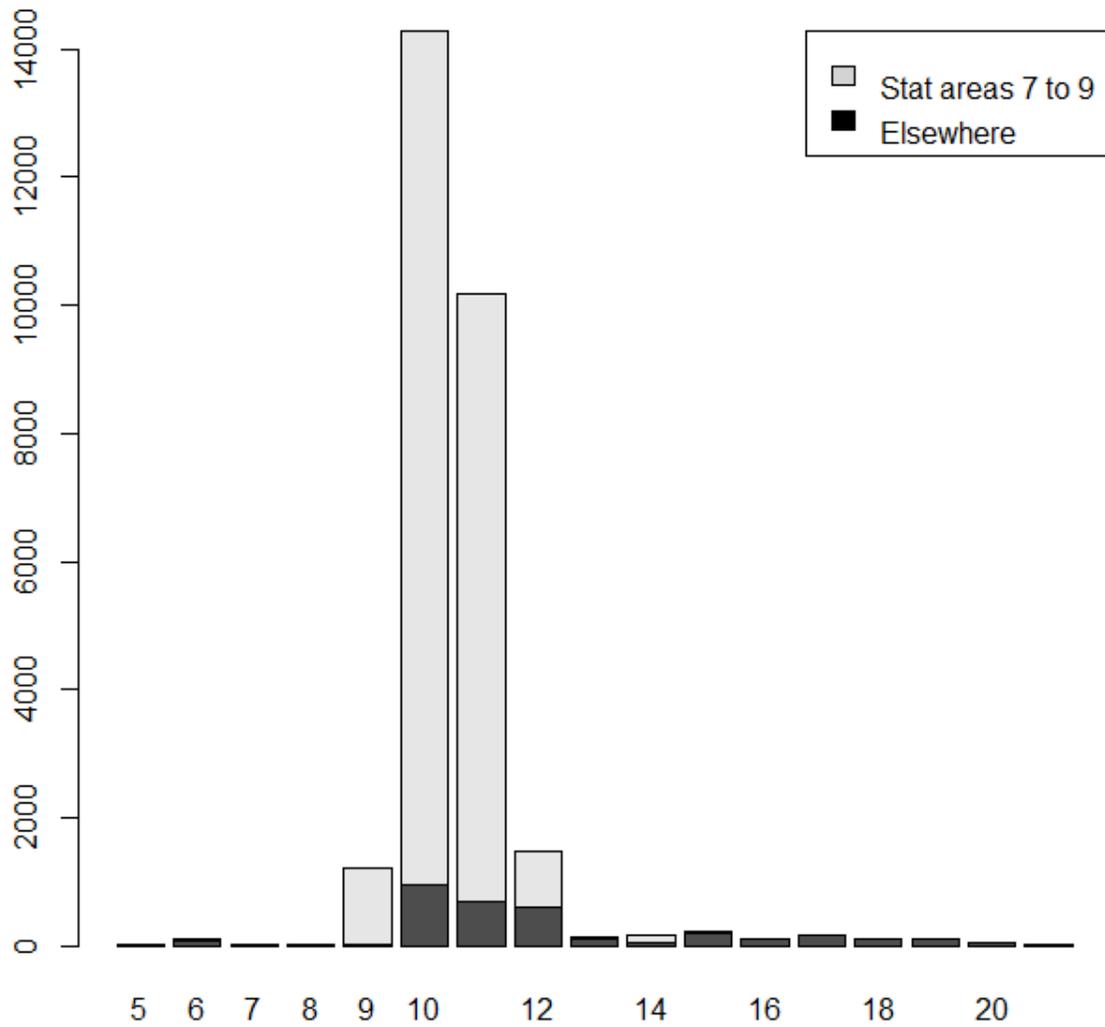


Figure 4: Frequency table of HBF for the main fishing ground with the lighter shade for statistical areas 7-9, and the darker shade for other areas.

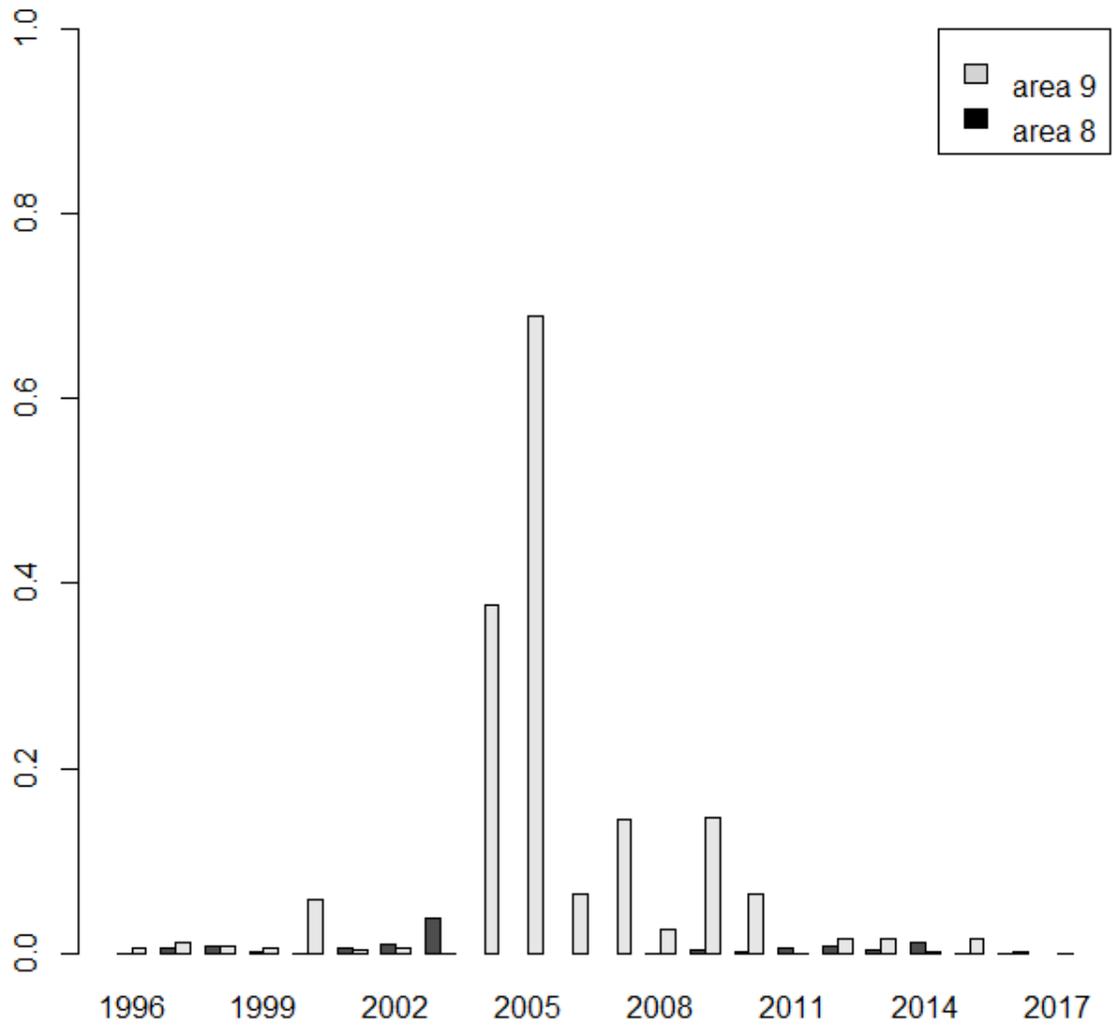


Figure 5: Proportions of sets with zero catches of SBT by year and statistical area, in the data used in the standardization models.

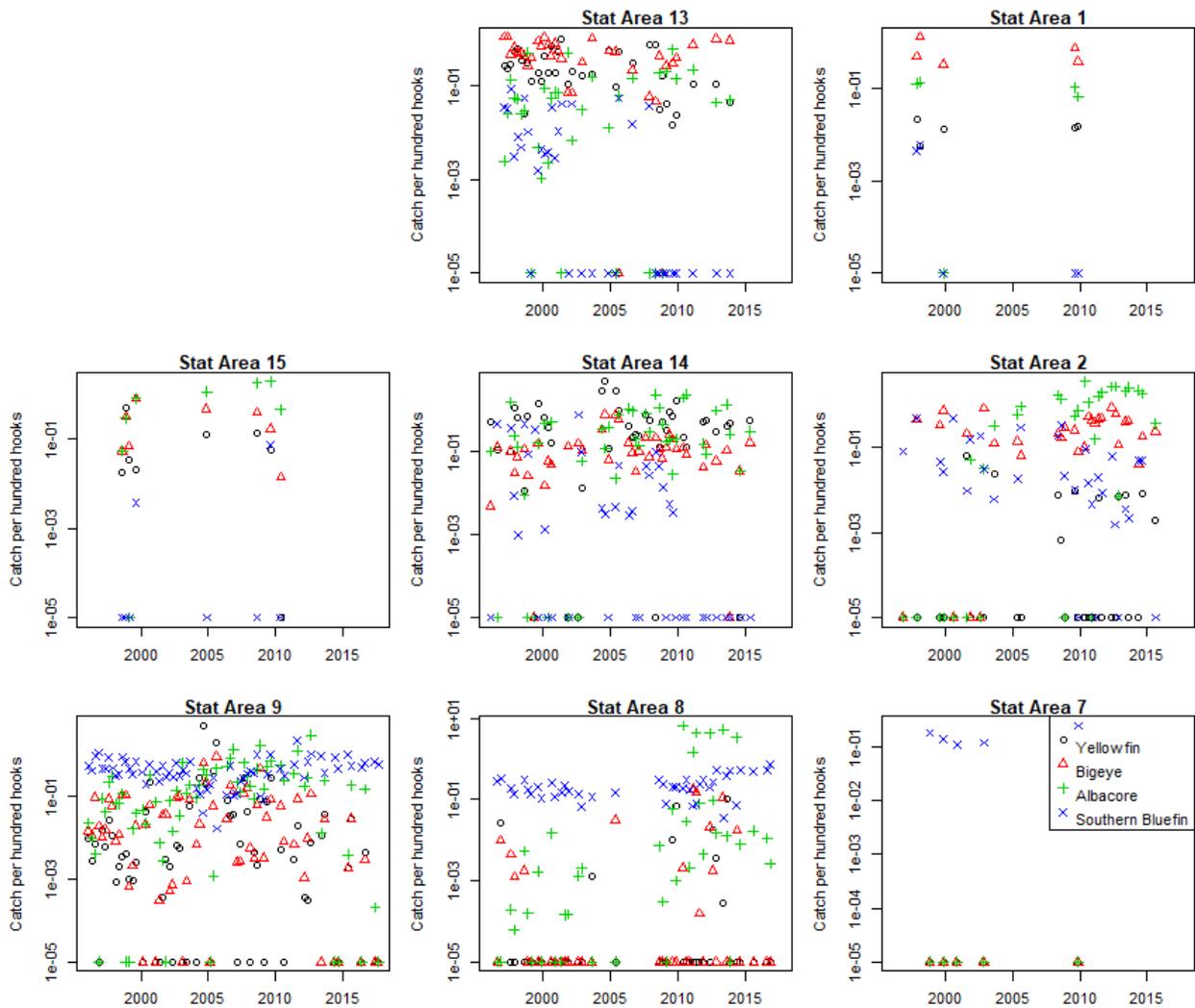


Figure 6: Mean catch per hundred hooks by year-qttr, species, and statistical area, plotted on a log scale, for yellowfin, bigeye, albacore, and southern bluefin tuna. Each CPUE has 1E-5 added so that zero catches appear on the log scale.

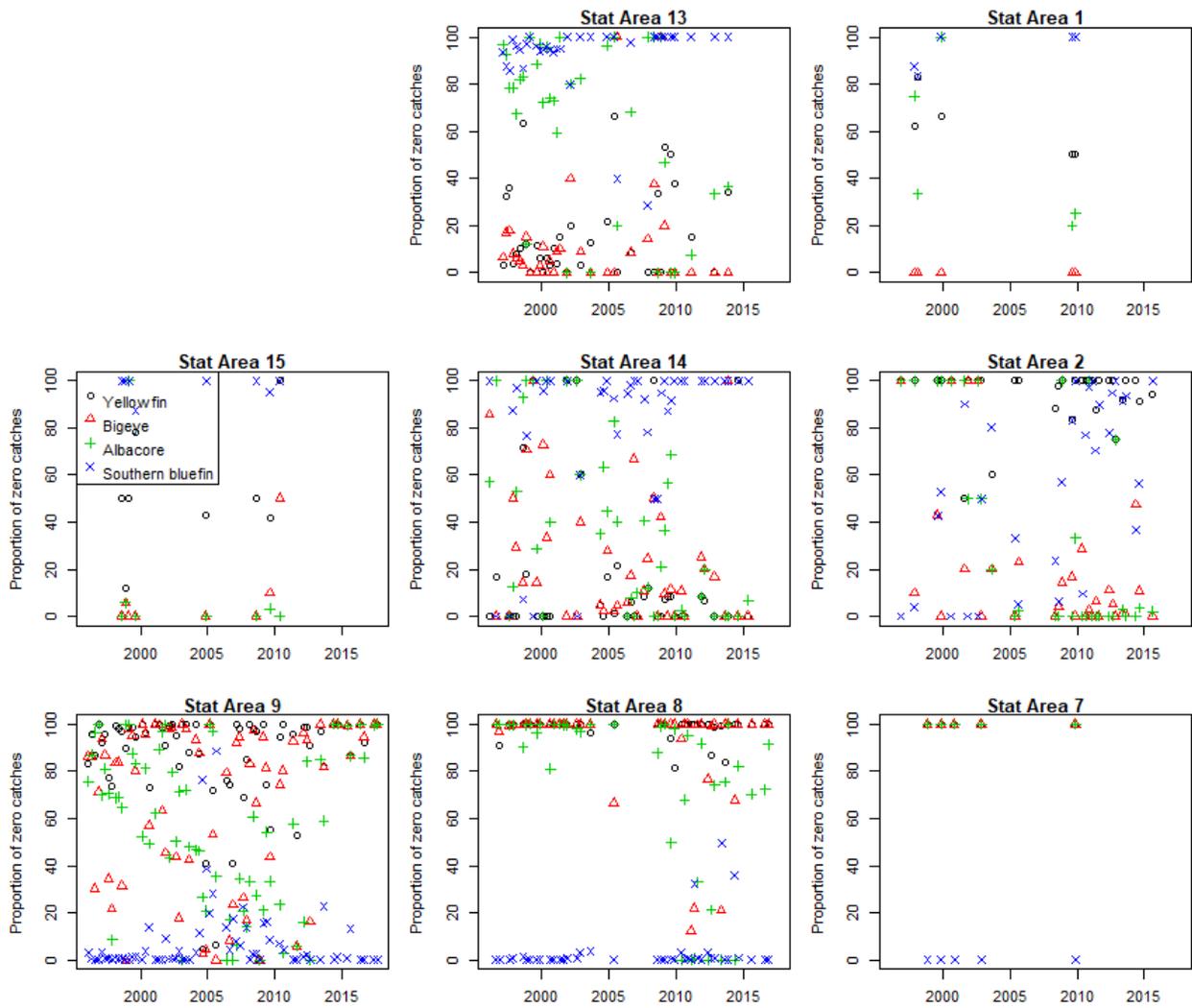


Figure 7: Proportion of zero catches per set by year-qtr, species, and statistical area, for yellowfin, bigeye, albacore, and southern bluefin tuna.

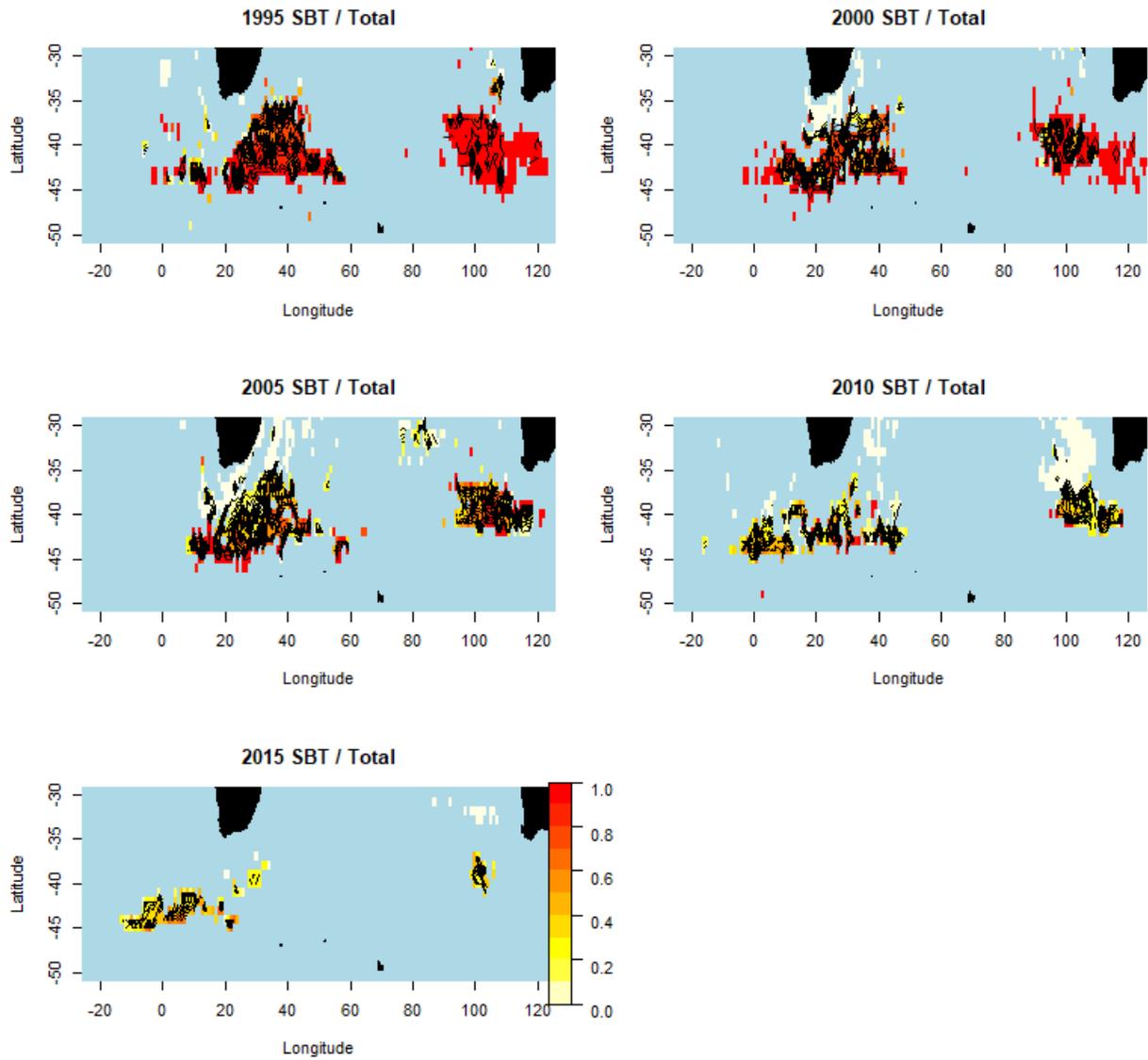


Figure 8: Proportion southern bluefin tuna (SBT) in the total reported catch in numbers by 1° cell, aggregated over 5 years within the period 1990-2014. Red colour indicates a higher proportion of SBT.

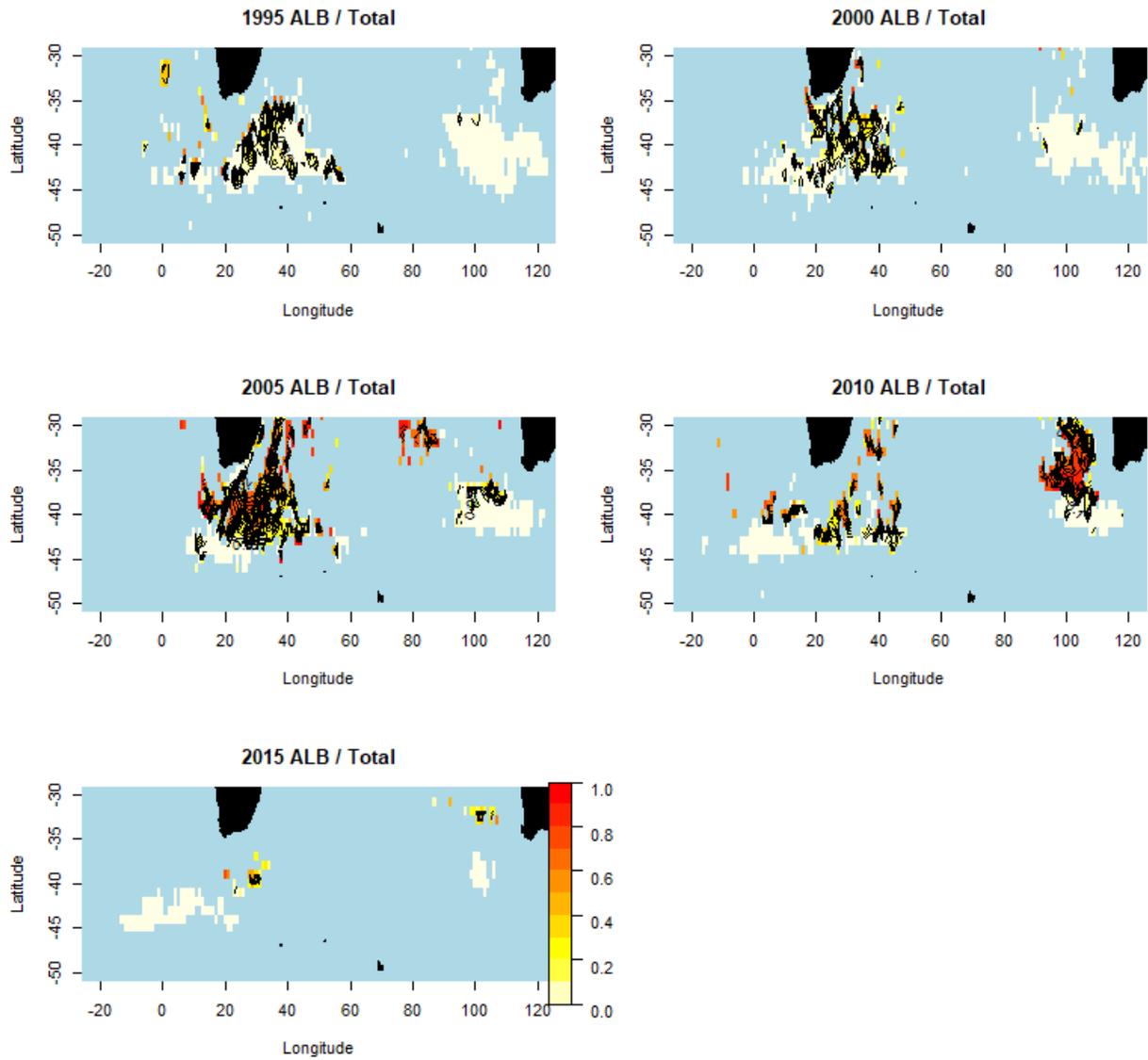


Figure 9: Proportion albacore (ALB) in the total reported catch in numbers by 1° cell, aggregated over 5 years within the period 1990-2014. Red colour indicates a higher proportion of ALB.

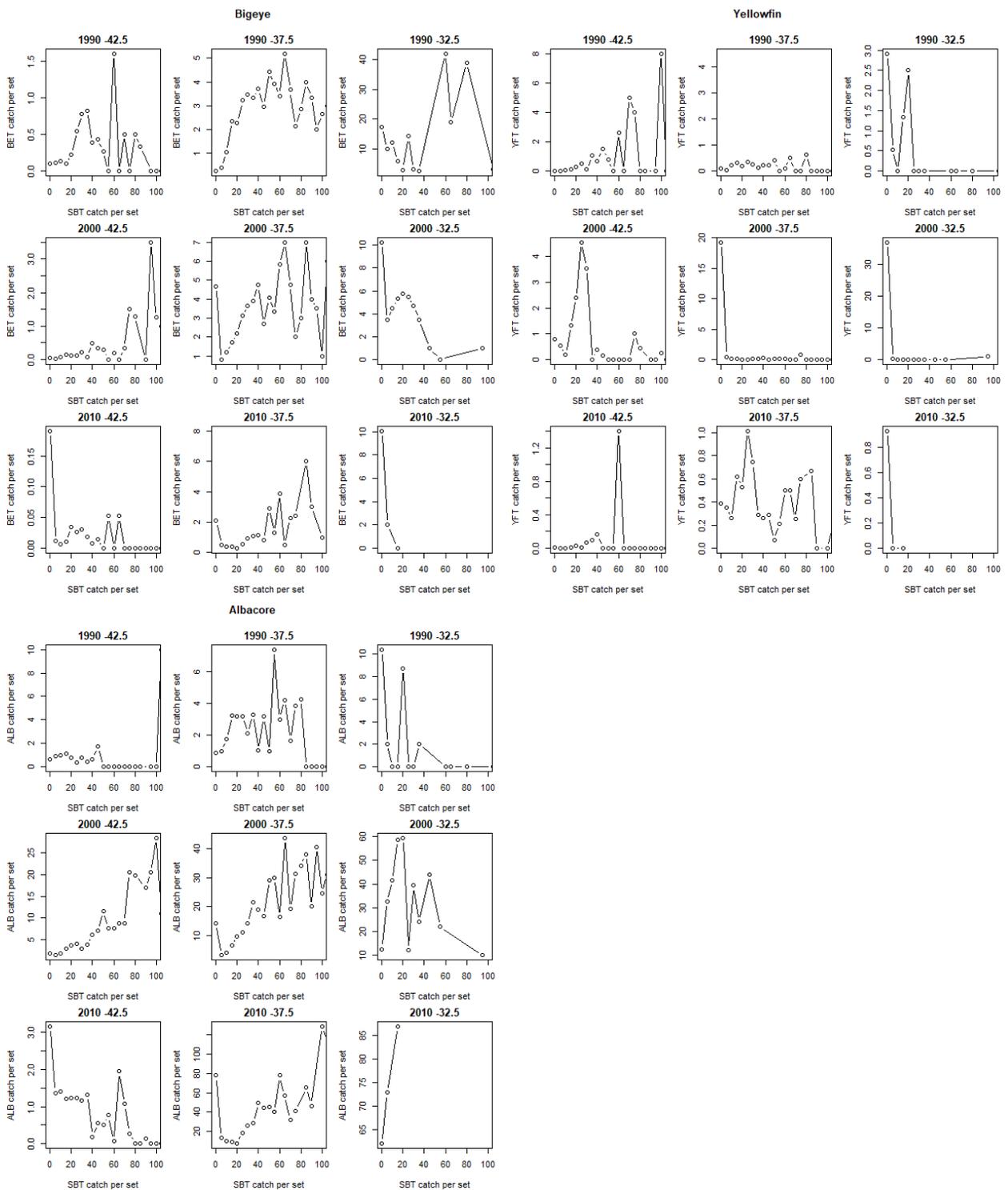


Figure 10: Mean catch per set of other species versus catch per set of southern bluefin tuna. SBT catches are binned by 5 fish. Plots are grouped by decade (start year) and latitude.

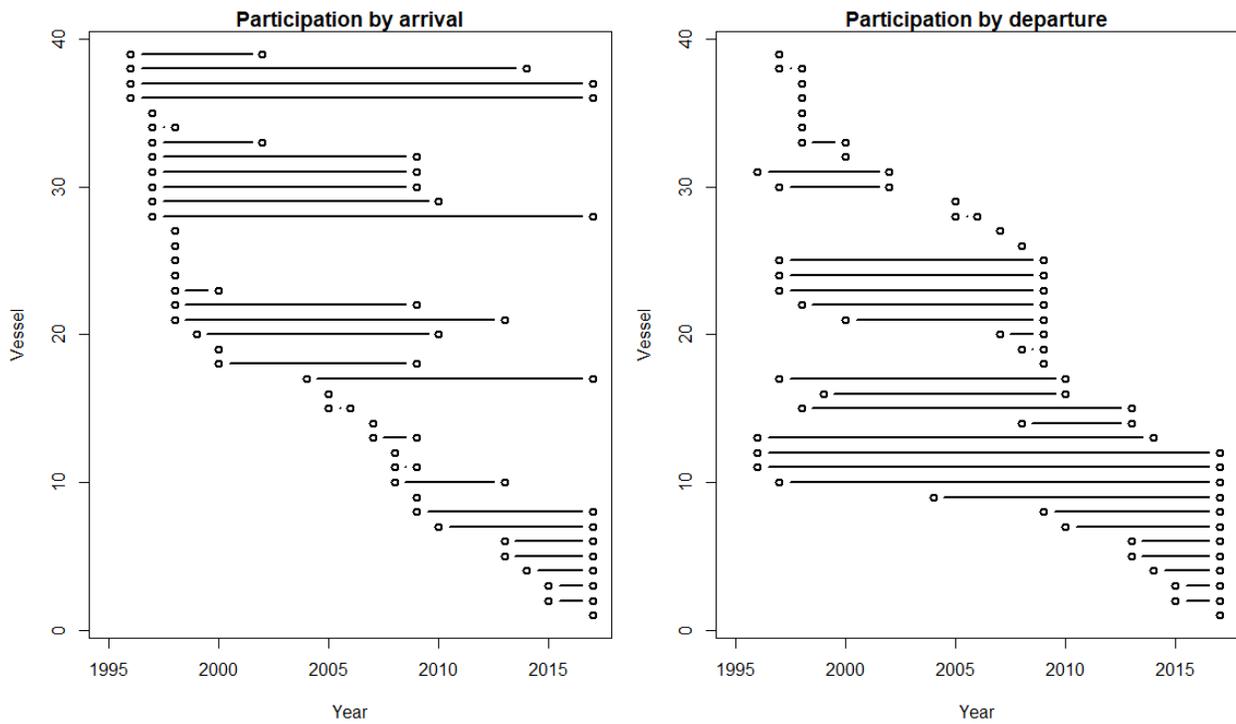


Figure 11: Plots of participation by vessel and year. Each row represents a vessel, with the left plot sorted by the first year of participation, and the right plot sorted by the final year.

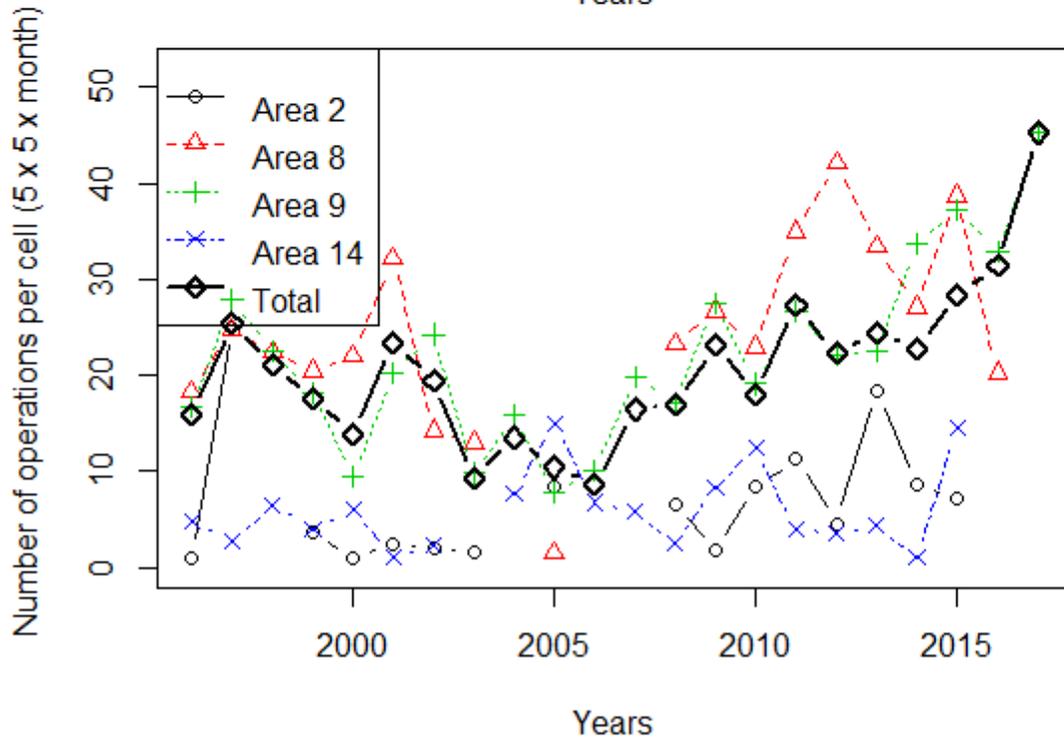
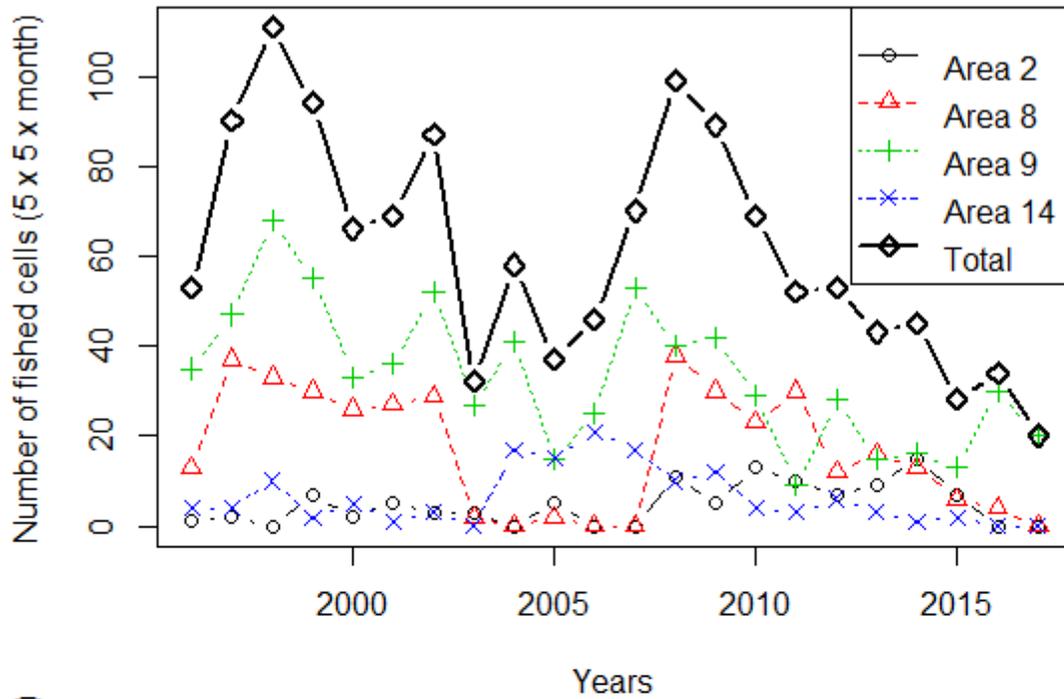


Figure 12: For fishing since 1996 in areas 2, 8, 9, and 14, the number of cells (5° latitude by 5° longitude by month) fished (above) and the number of longline operations per cell (below).

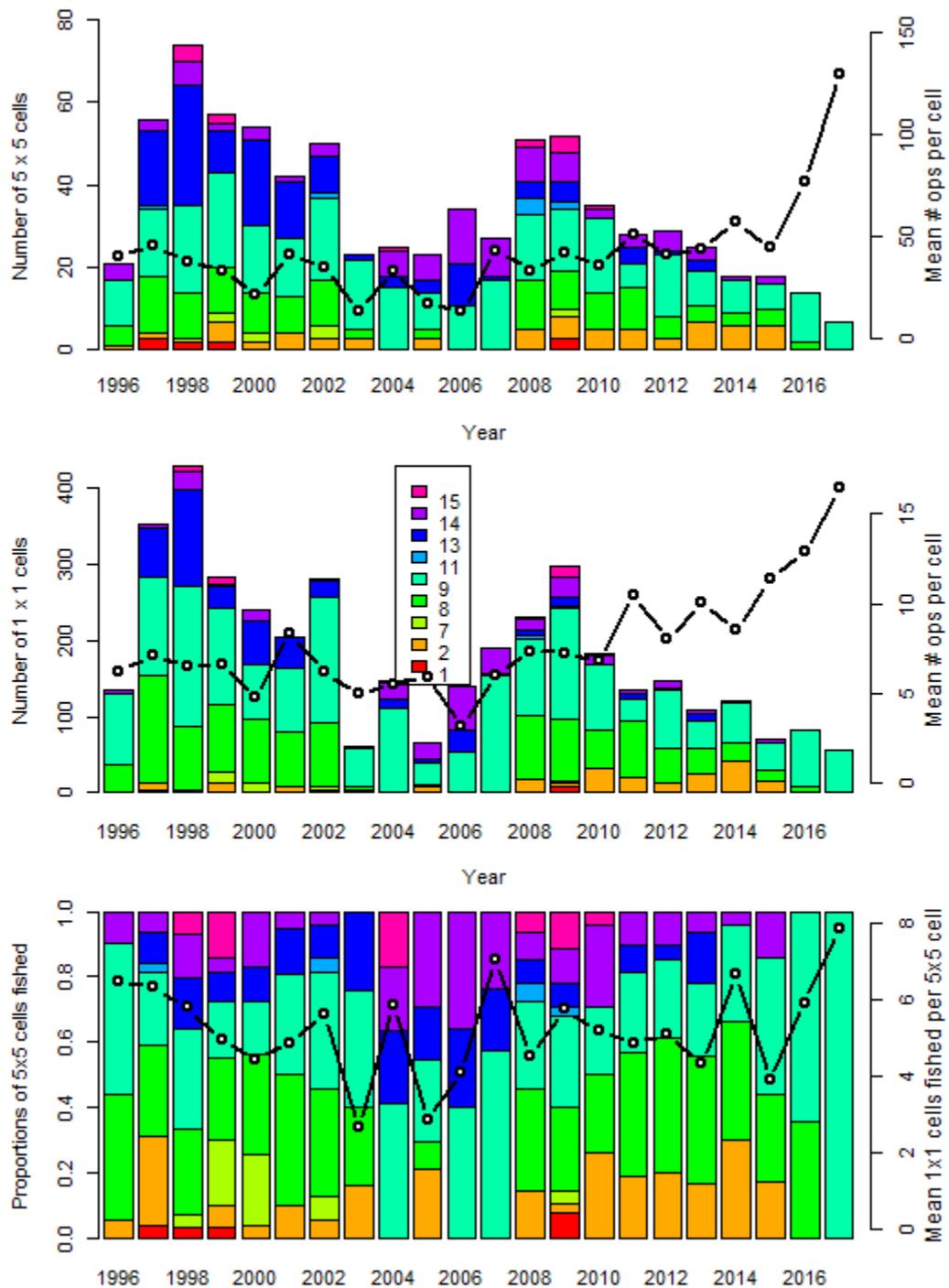


Figure 13: (Upper) Bars represent the number of major cells (5x5° by month) fished by CCSBT statistical area and year, see left y-axis. The line represents the mean annual operations per cell, see right y-axis. (Middle) As for upper plot, but with minor cells (1x1° by month) instead of major cells. (Lower) Relative distribution of fished major cells by the proportion of the cell fished, measured as the number of minor cells fished within each major cell (see left y-axis). The lowest (red) and highest (purple) bands represent major cells in which, respectively, 1 and 15 of the 25 minor cells were fished. The line represents the mean number of minor cells fished per major cell by year, see right y-axis.

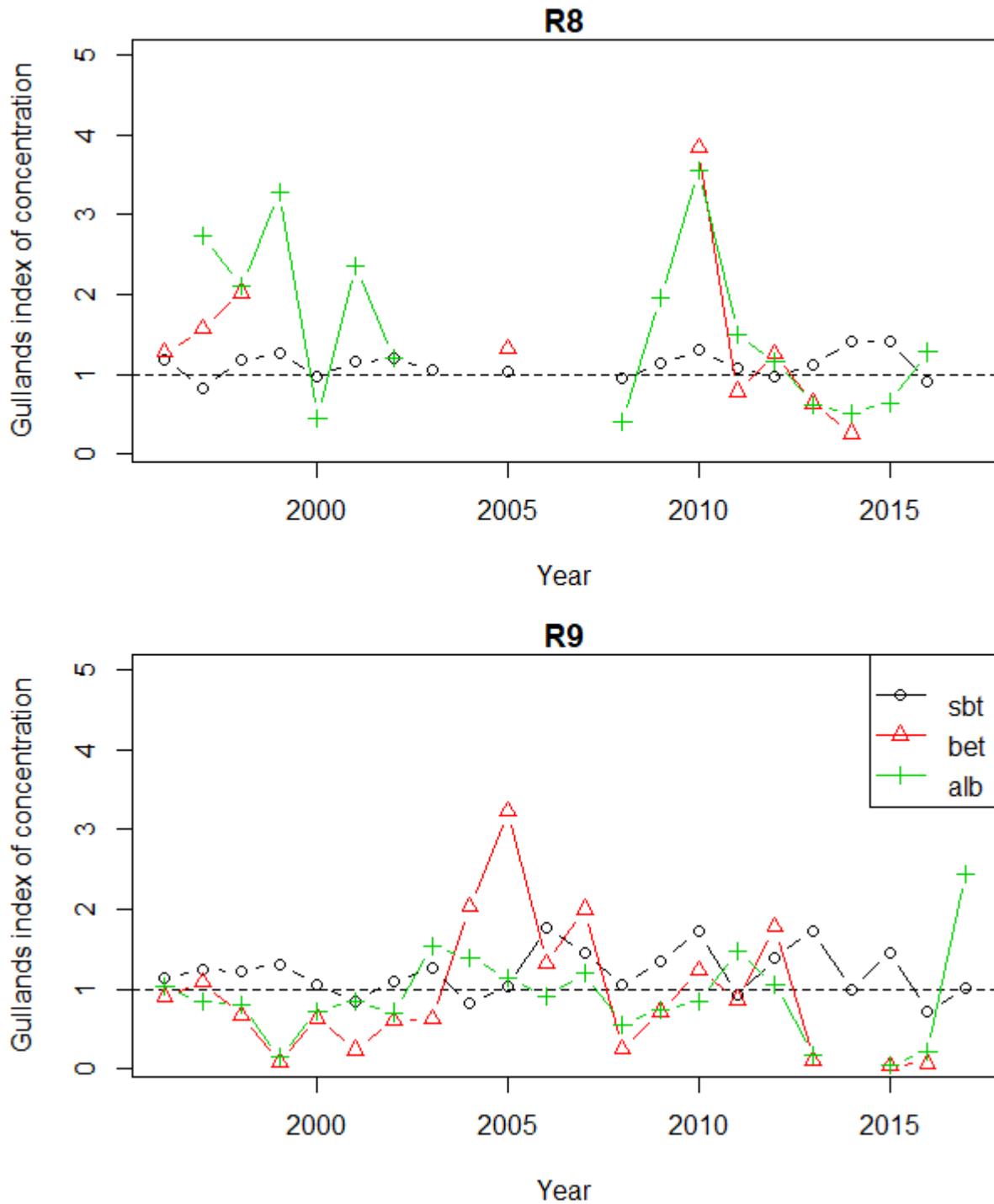


Figure 14: Gulland's indices of concentration estimated annually for southern bluefin tuna, bigeye tuna, and albacore tuna, in statistical areas 8 and 9.

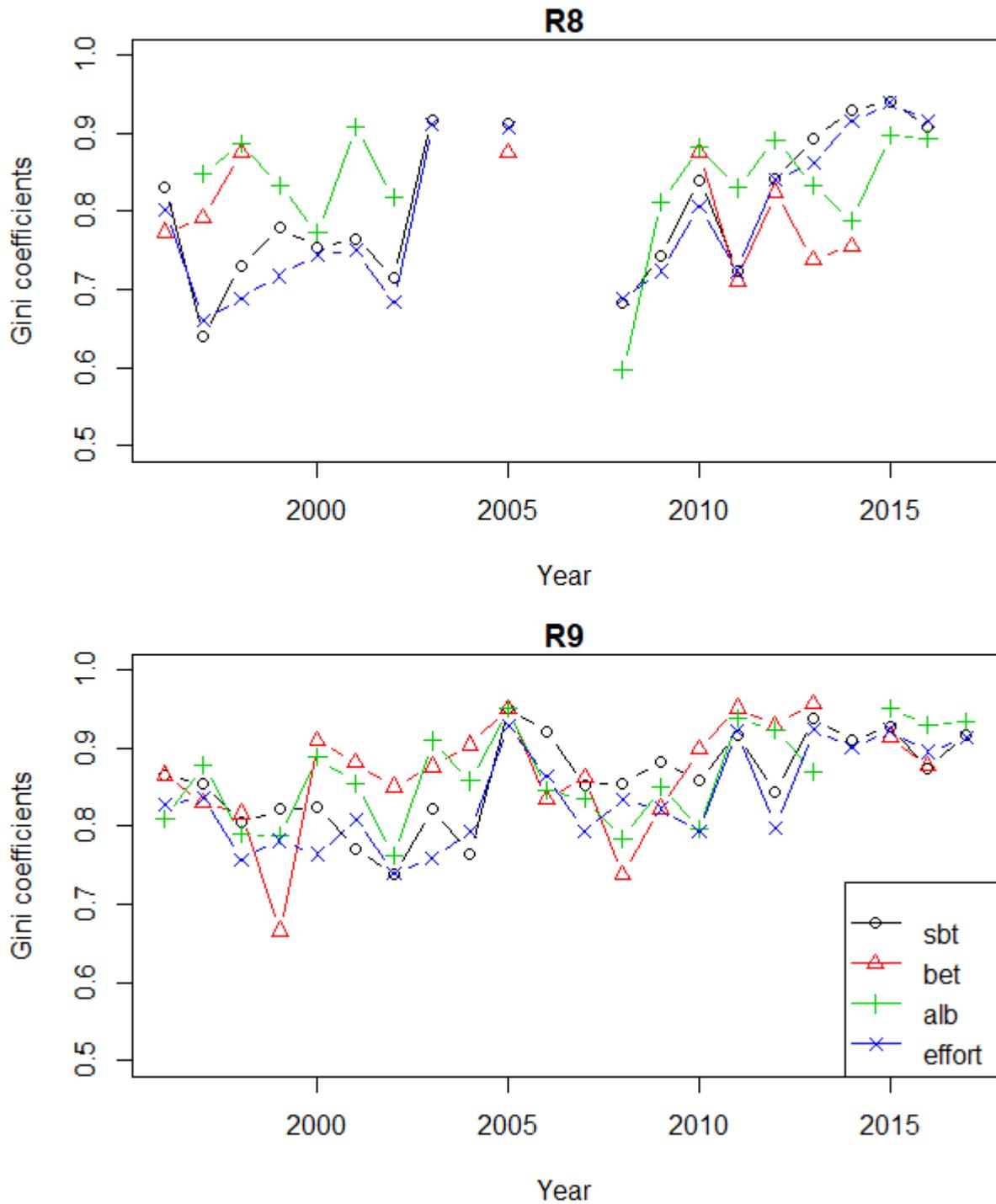
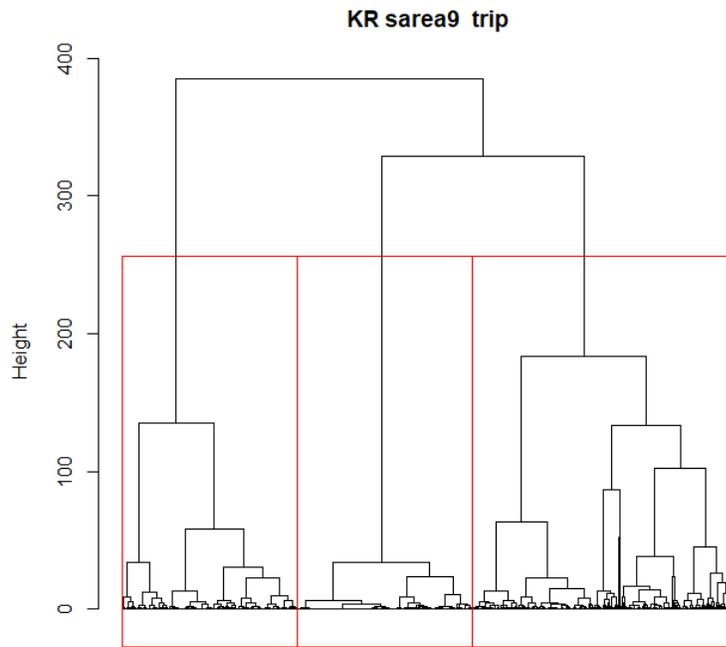
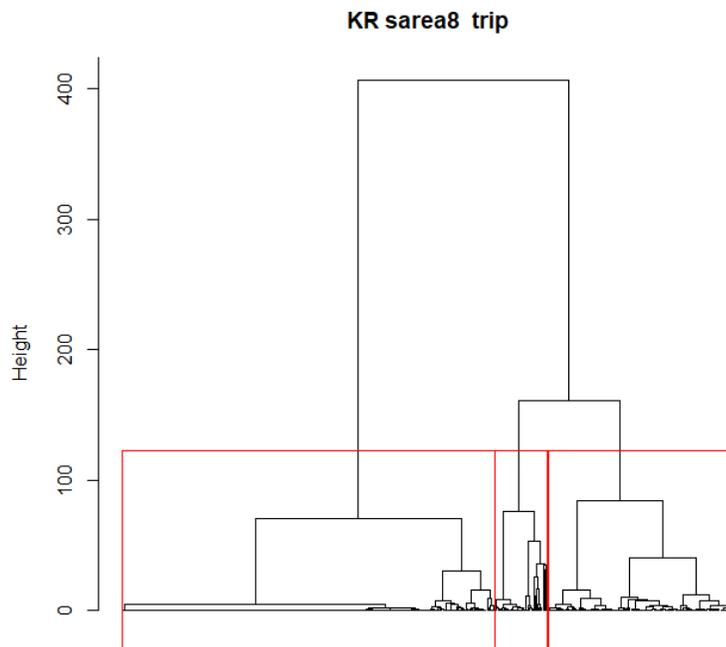


Figure 15: Gini coefficients estimated annually for southern bluefin tuna, bigeye tuna, albacore tuna, and fishing effort in statistical areas 8 and 9



dtrip
hclust (*, "ward.D")



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Figure 16: Dendrograms for Ward hierarchical cluster analyses of statistical areas 9 (above) and 8 (below), with the red lines indicating the separation into 3 clusters for each.

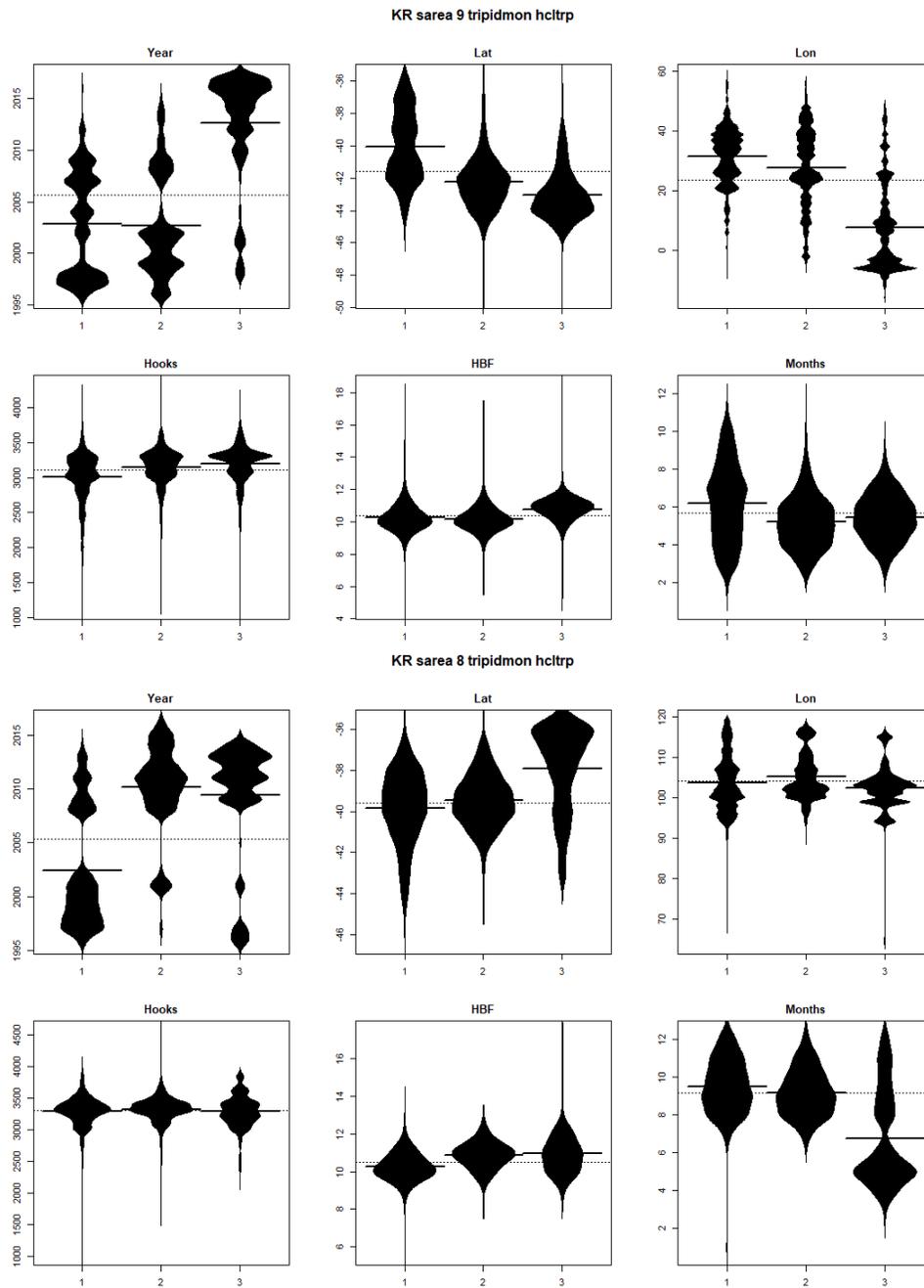


Figure 17: Beanplots for statistical areas 9 (above) and 8 (below), showing the number of sets versus covariate by cluster. The horizontal bars indicate the medians.

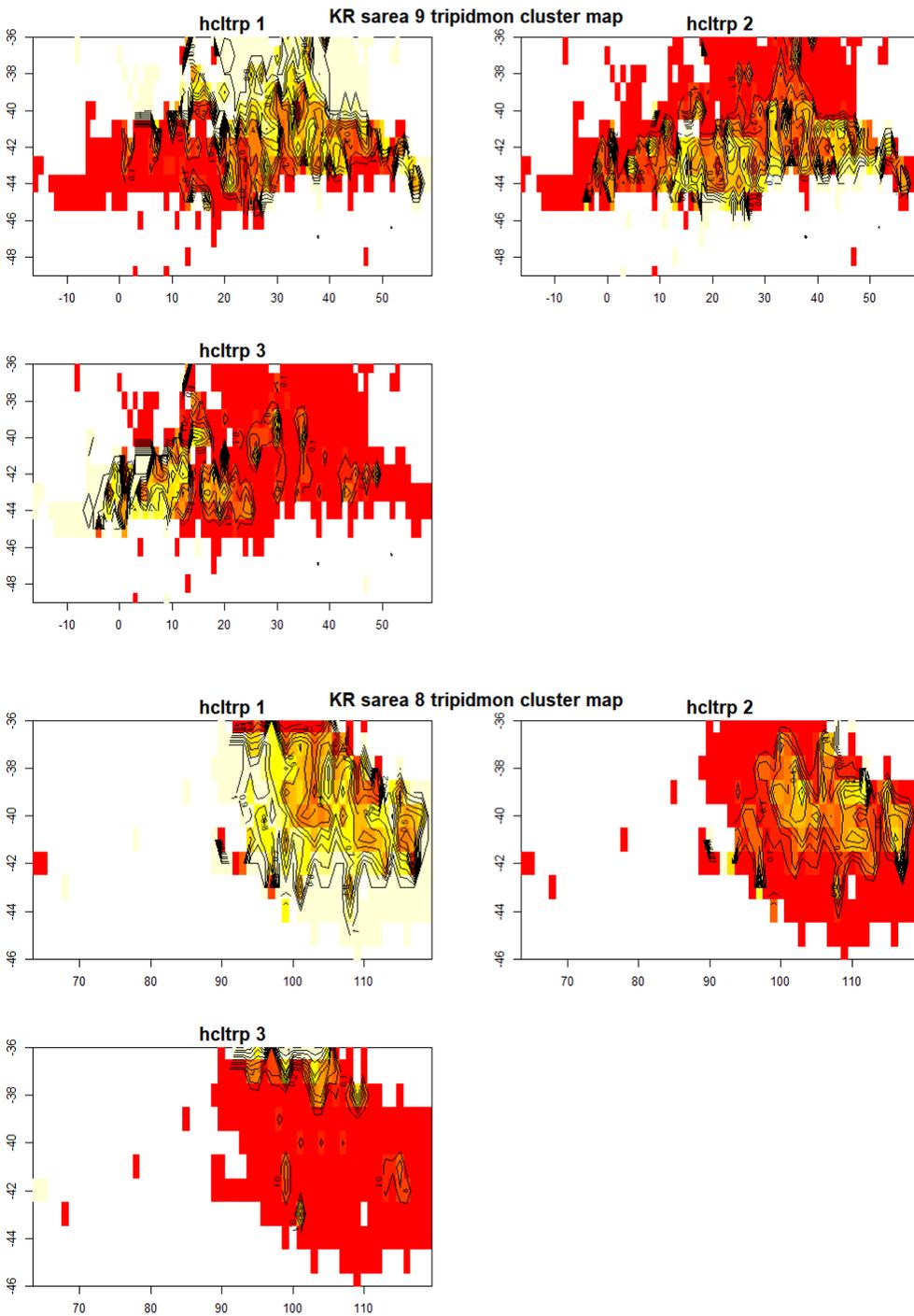


Figure 18: Maps of the proportion of each cluster per 1 degree square in total effort for statistical areas 9 (above) and 8 (below). Higher proportions are shown in yellow. White space indicates no reported effort.

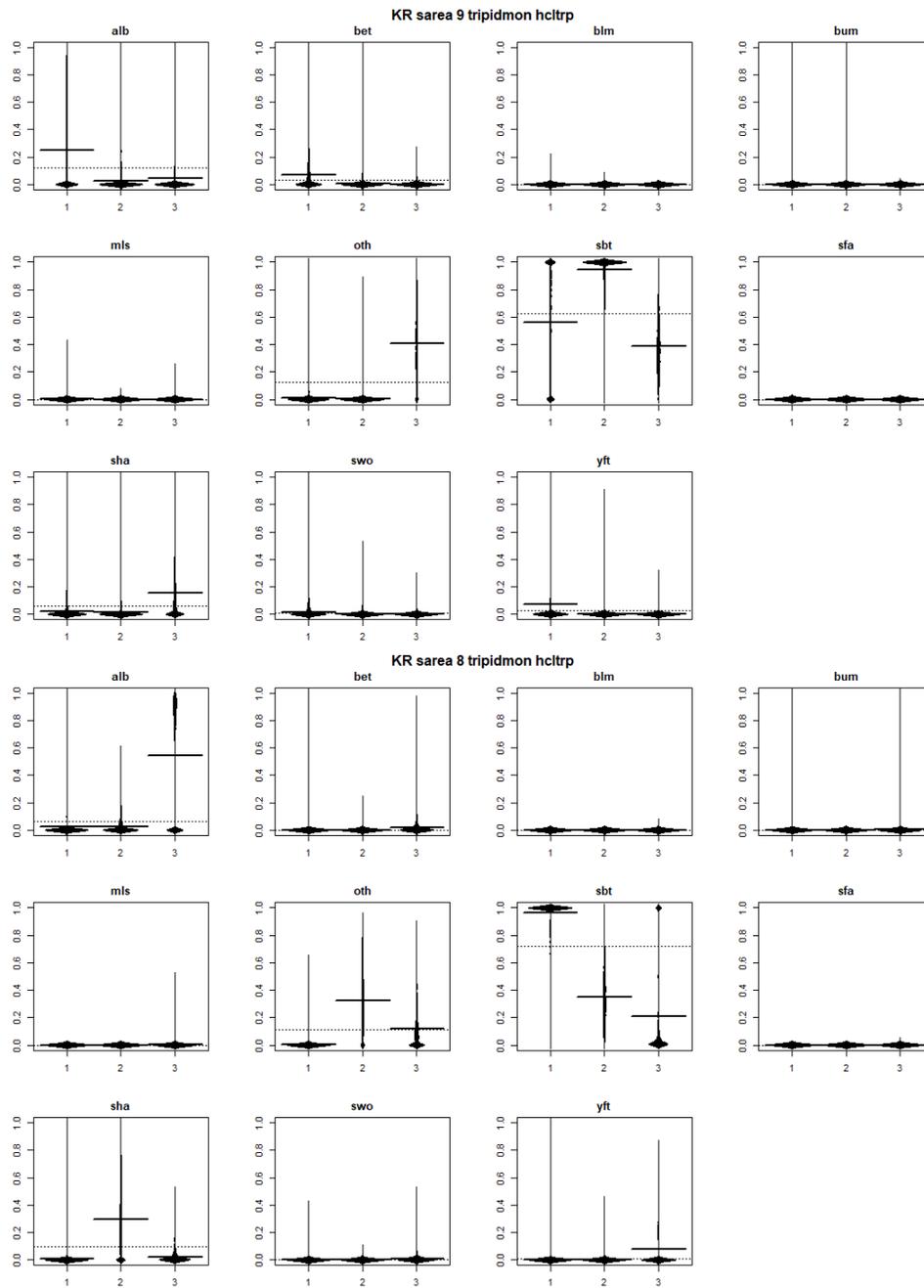


Figure 19: Beanplots for statistical areas 9 (above) and 8 (below), showing species composition by cluster. The horizontal bars indicate the medians.

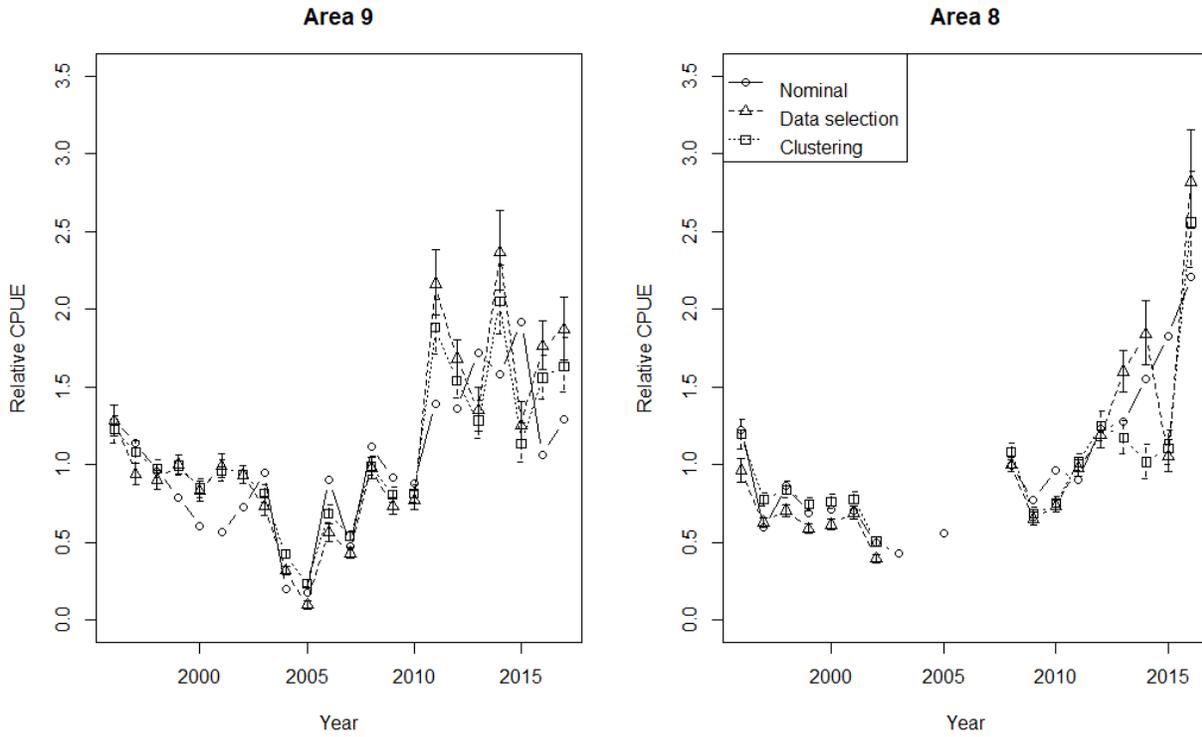


Figure 20: Nominal and standardized CPUE indices based on lognormal GLMs with an added constant, addressing target change using selected data (triangles) and cluster analysis (squares), for statistical areas 9 (left) and 8 (right).

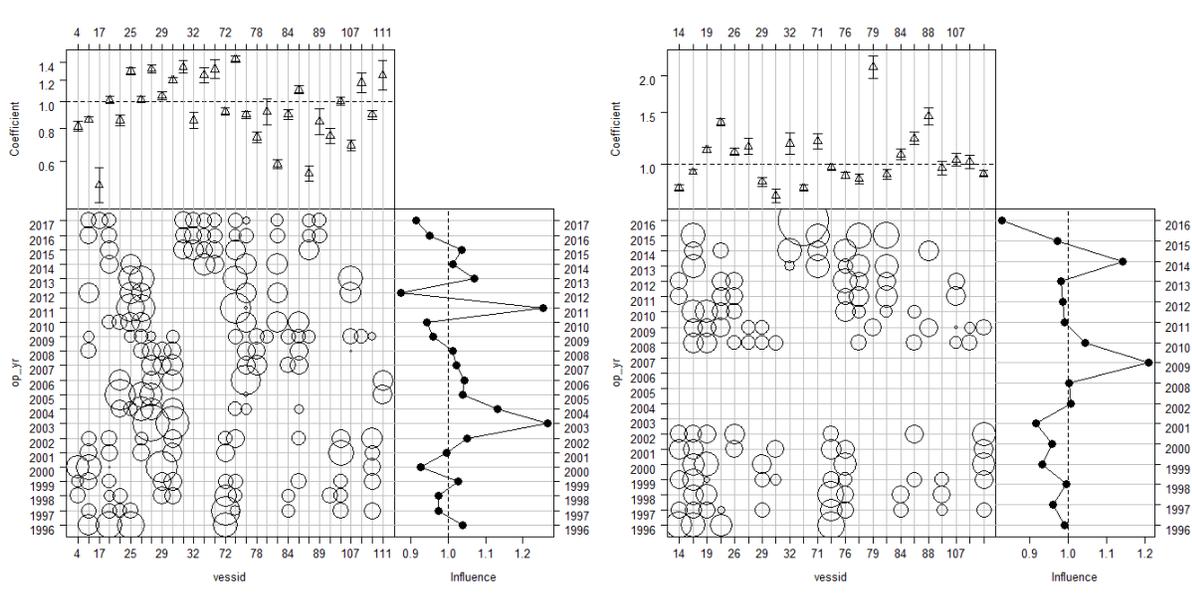


Figure 21: Influence plots for vessel effects for areas 9 (left) and 8 (right), addressing target change using clustering.

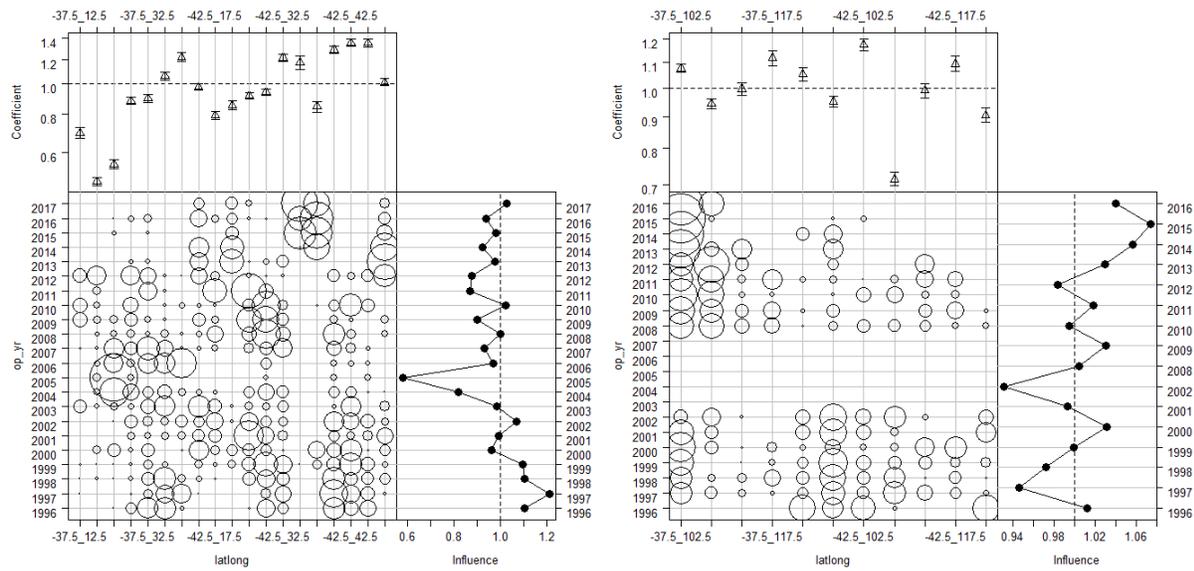


Figure 22: Influence plots for spatial latlong effects for statistical areas 9 (left) and 8 (right) , addressing target change using clustering.

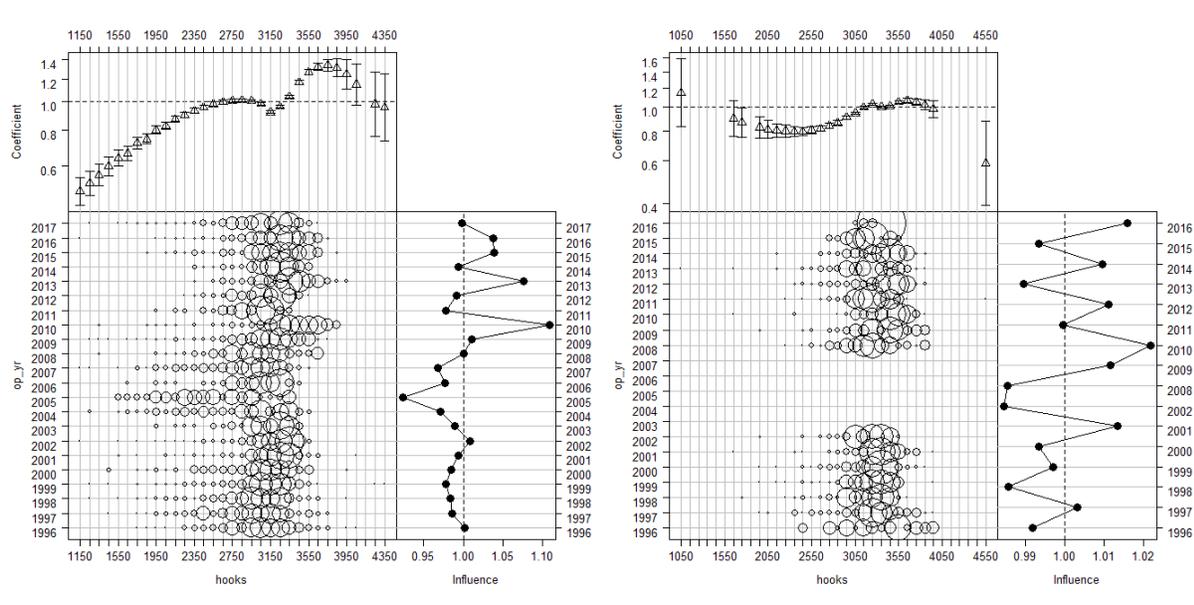


Figure 23: Influence plots for the effects of numbers of hooks for statistical areas 9 (left) and 8 (right), addressing target change using clustering.

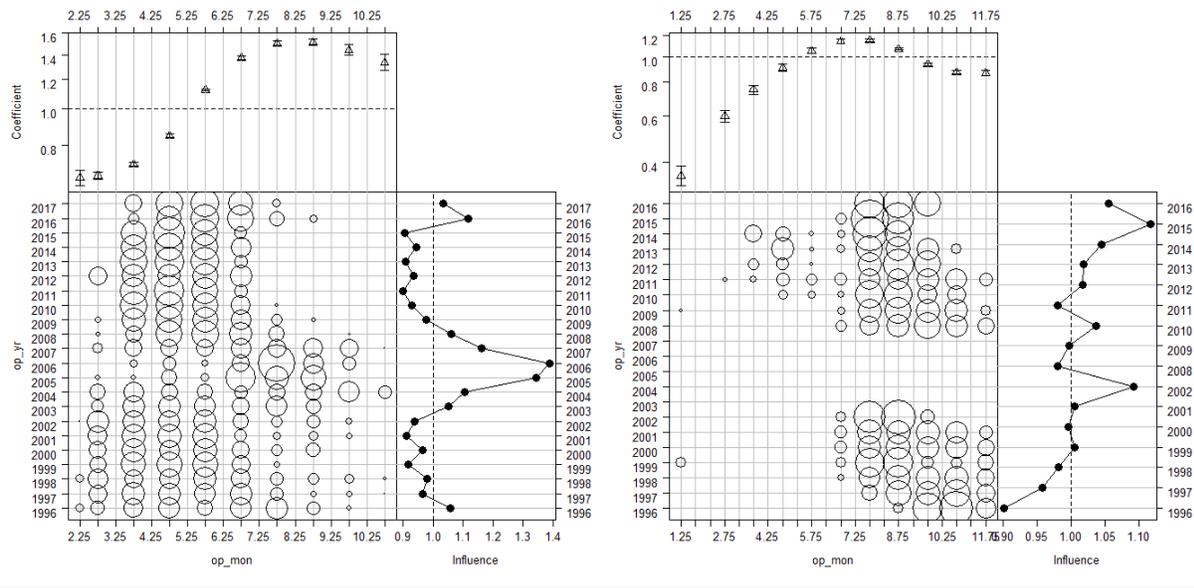


Figure 24: Influence plots for month effects for statistical areas 9 (left) and 8 (right), addressing target change using clustering.

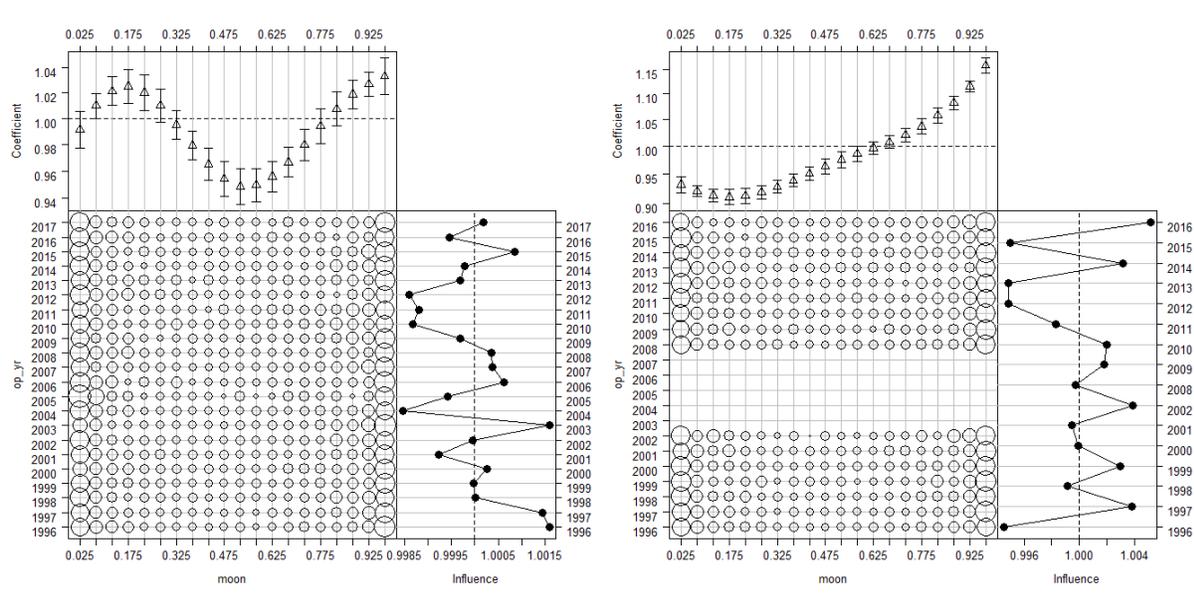


Figure 25: Influence plots for lunar illumination effects for statistical areas 9 (left) and 8 (right), addressing target change using clustering.

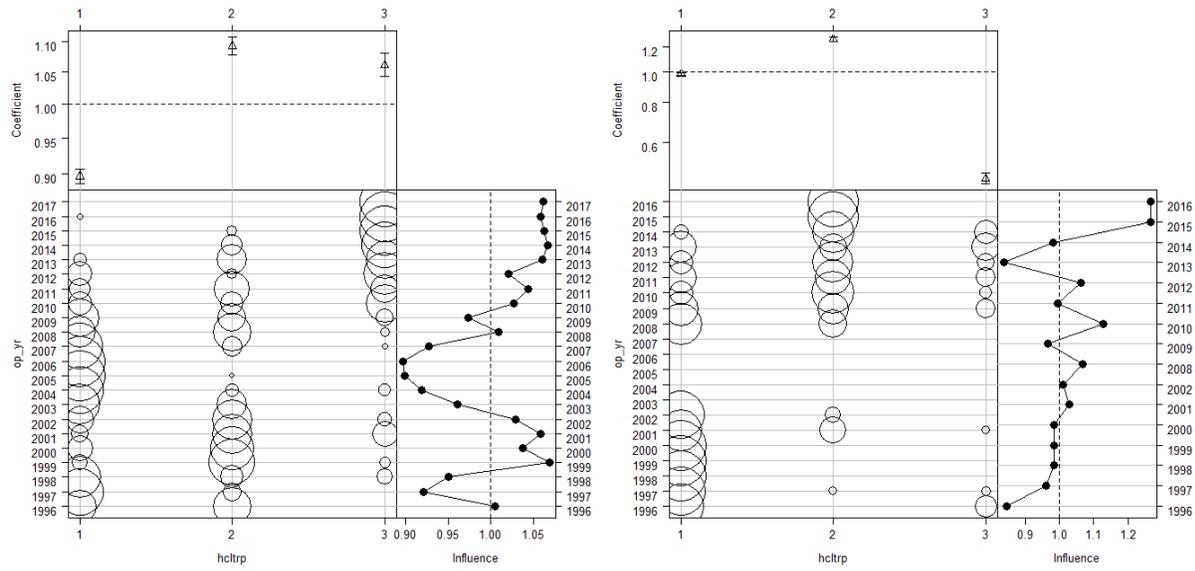


Figure 26: Influence plots for cluster effects for statistical areas 9 (left) and 8 (right), addressing target change using clustering.

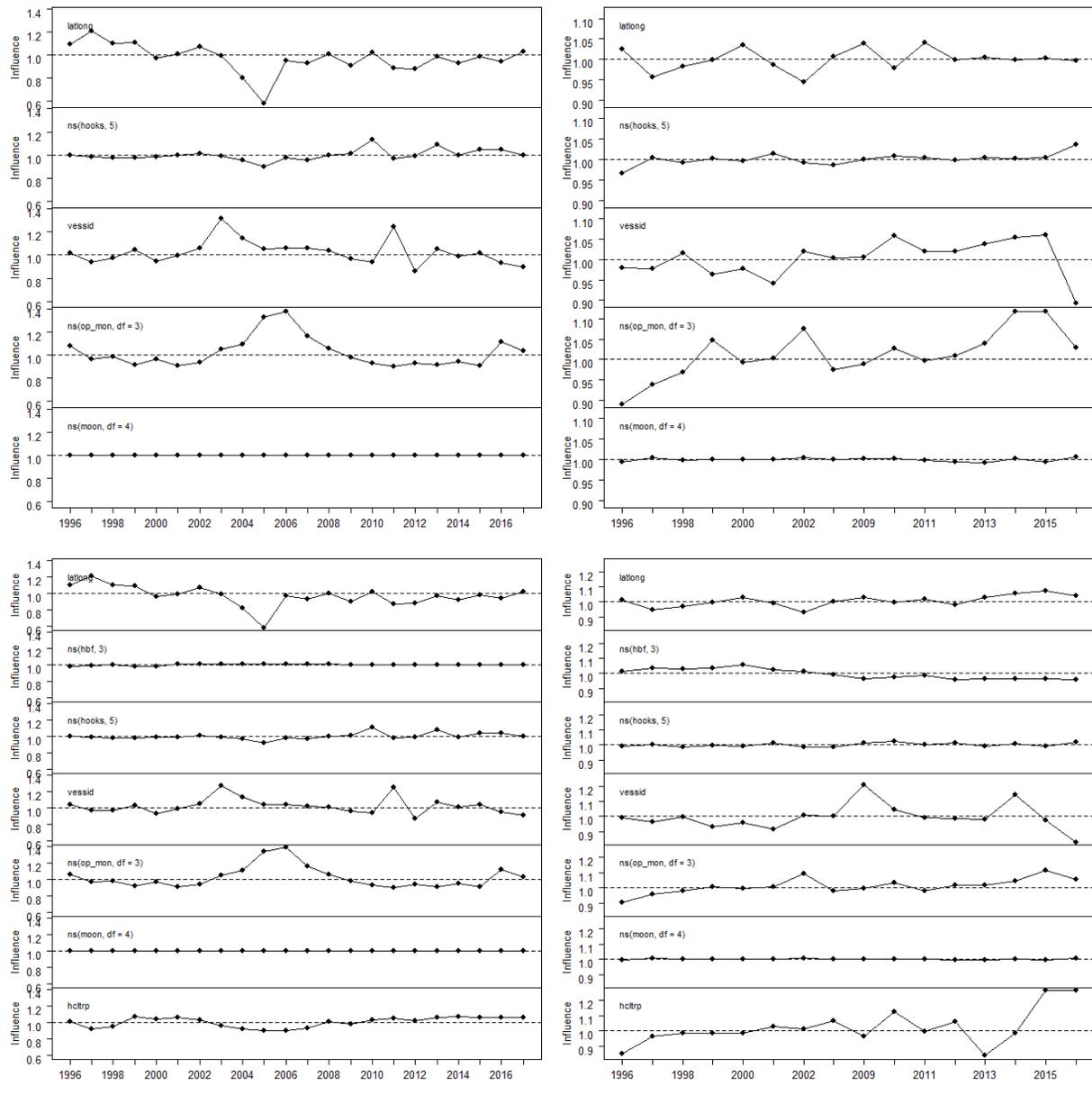


Figure 27: Compilation of influence plots for statistical areas 9 (left) and 8 (right), addressing target change using selected data (above) and clustering (below).

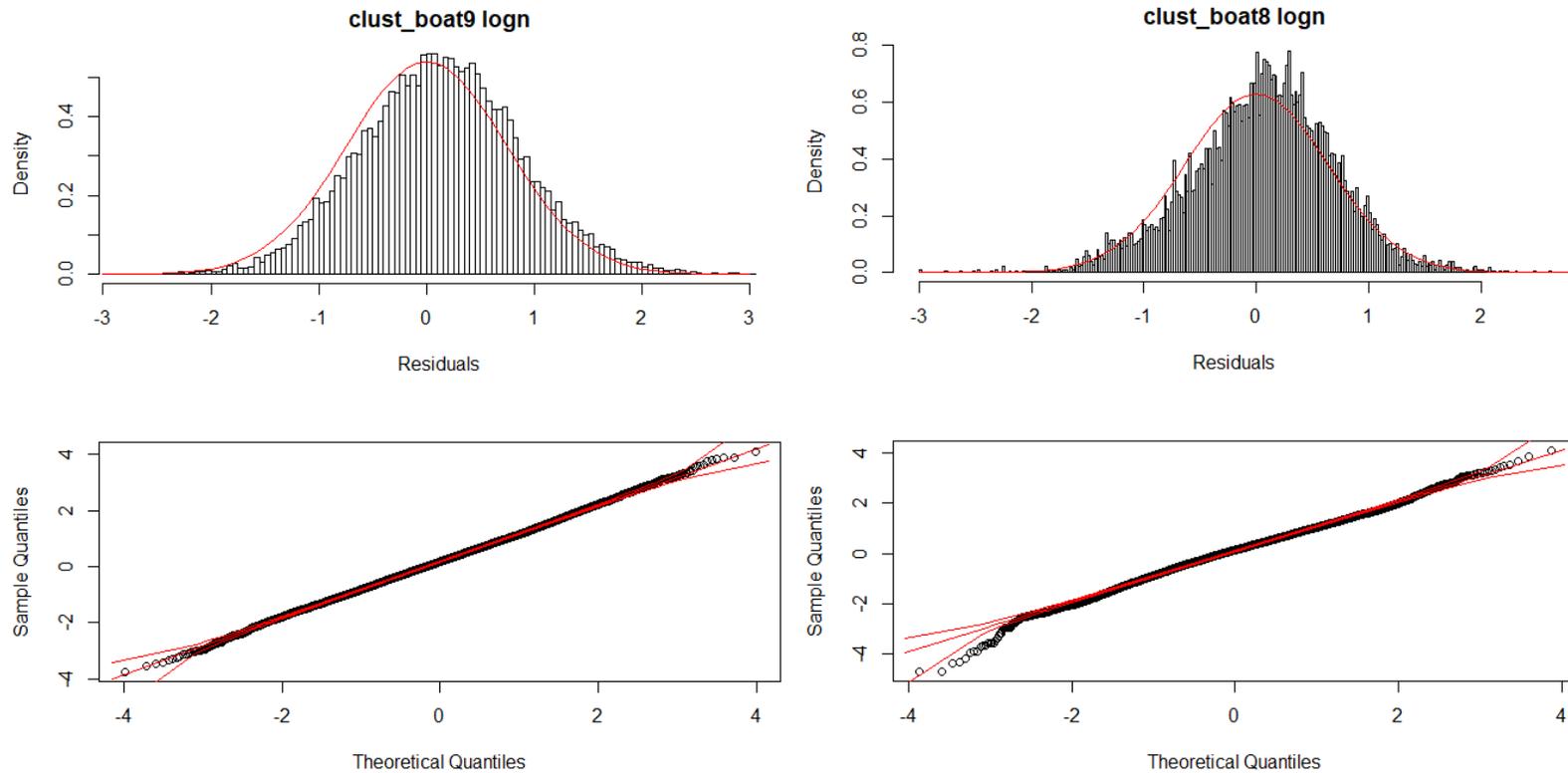


Figure 28: Frequency distributions of the standardized residuals (above) and Q-Q plots of standardized residuals for lognormal constant GLM analyses of statistical areas 9 (left) and 8 (right), based on the model with cluster as a covariate.

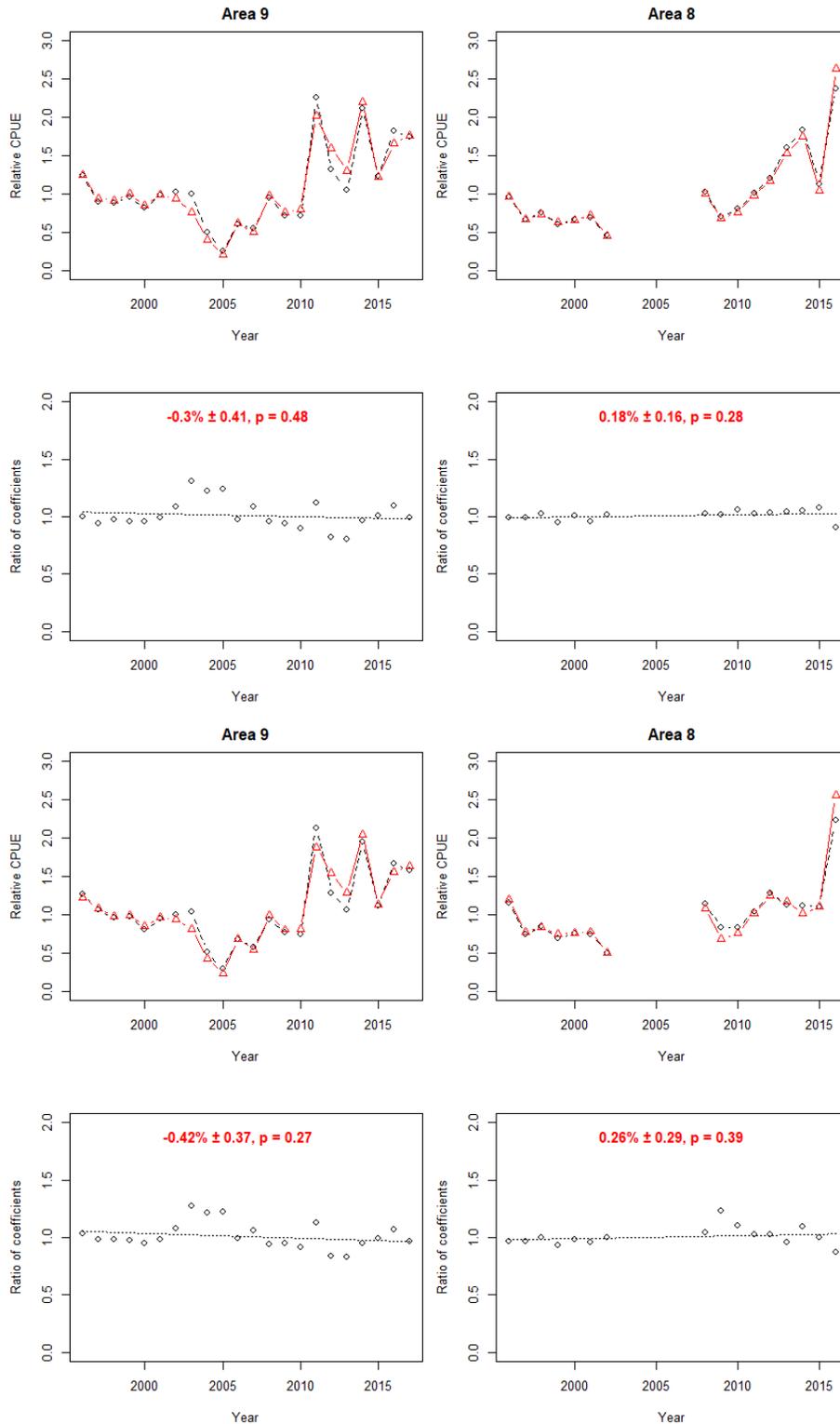


Figure 29: Annual abundance indices from standardizing SBT CPUE for statistical areas 9 (left) and 8 (right) using lognormal constant models, fitted either with (red triangles) or without (black circles) vessel effects. The second and fourth rows show the ratios of each pair of indices, with log-linear trends fitted. The numbers indicate the annual rate of change in the ratio. The top four plots use selected data, while the lower four plots use clustered data.